

US EPA ARCHIVE DOCUMENT

Final

Total Maximum Daily Loads

for

Dissolved Oxygen & Nutrients in

Pinellas Park Ditch No. 1

WBID 1662

July 2013



In compliance with the provisions of the Federal Clean Water Act, 33 U.S.C §1251 et. seq., as amended by the Water Quality Act of 1987, P.L. 400-4, the U.S. Environmental Protection Agency is hereby establishing the Total Maximum Daily Load (TMDL) for dissolved oxygen and nutrients in the Springs Coast Basin (WBID 1662). Subsequent actions must be consistent with this TMDL.

/s/

James D. Giattina, Director
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7/7/2013

Date

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SUMMARY SHEET for WBID 1662

Total Maximum Daily Load (TMDL)

2006 303(d) Listed Waterbodies for TMDLs addressed in this report:

WBID	Segment Name	Class and Waterbody Type	Major River Basin	HUC	County	State
1662	Pinellas Park Ditch No 1 (Tidal Segment)	Class III Marine	Springs Coast	3100207	Pinellas	Florida

TMDL Endpoints/Targets:

Dissolved Oxygen & Nutrients

TMDL Technical Approach:

The TMDL allocations were determined by analyzing the effects of TN, TP, and BOD concentrations on DO concentrations in the waterbody. A watershed model and estuary model were used to predict delivery of pollutant loads to the waterbody and to evaluate the in-stream impacts of the pollutant loads.

TMDL Waste Load and Load Allocation

Constituent	Current Condition		TMDL Condition		Percent Reduction		
	WLA (kg/yr)	LA (kg/yr)	WLA (kg/yr)	LA (kg/yr)	WLA	LA	MS4
Total Nitrogen	--	5,905	--	1,200	--	80%	80%
Total Phosphorus	--	697	--	300	--	85%	85%
Biochemical Oxygen Demand	--	16,620	--	5,422	--	67%	67%

Endangered Species Present (Yes or Blank):

USEPA Lead TMDL (USEPA or Blank): USEPA

TMDL Considers Point Source, Non-point Source, or Both: Both

Major NPDES Discharges to surface waters addressed in USEPA TMDL:

Permit ID	Permittee(s)	County	Permit Type
FLS000005	Pinellas County, City of Pinellas Park, Florida Department of Transportation	Pinellas	Phase I MS4

1.0 INTRODUCTION

Section 303(d) of the Clean Water Act requires each state to list those waters within its boundaries for which technology based effluent limitations are not stringent enough to protect any water quality standard applicable to such waters. Listed waters are prioritized with respect to designated use classifications and the severity of pollution. In accordance with this prioritization, states are required to develop Total Maximum Daily Loads (TMDLs) for those water bodies that are not meeting water quality standards. The TMDL process establishes the allowable loadings of pollutants or other quantifiable parameters for a waterbody based on the relationship between pollution sources and in-stream water quality conditions, so that states can establish water quality based controls to reduce pollution from both point and nonpoint sources and restore and maintain the quality of their water resources (USEPA 1991).

The Florida Department of Environmental Protection (FDEP) developed a statewide, watershed-based approach to water resource management. Under the watershed management approach, water resources are managed on the basis of natural boundaries, such as river basins, rather than political boundaries. The watershed management approach is the framework FDEP uses for implementing TMDLs. The state's 52 basins are divided into five groups and water quality is assessed in each group on a rotating five-year cycle. FDEP also established five water management districts (WMD) responsible for managing ground and surface water supplies in the counties encompassing the districts.

For the purpose of planning and management, the WMD divided the districts into planning units defined as either an individual primary tributary basin or a group of adjacent primary tributary basins with similar characteristics. These planning units contain smaller, hydrological based units called drainage basins, which are further divided by FDEP into "water segments". A water segment usually contains only one unique waterbody type (stream, lake, canal, etc.) and is about 5 square miles. Unique numbers or waterbody identification (WBIDs) numbers are assigned to each water segment. The WBID addressed in this report is located in the Springs Coast Basin and is part of the Anclote River/Coastal Pinellas County Planning Unit. The WBID is a Group 5 waterbody managed by the Southwest Florida Water Management District (SWFWMD). This TMDL report addresses WBID 1662, Pinellas Park Ditch No. 1, which is impaired for dissolved oxygen and nutrients.

2.0 PROBLEM DEFINITION

To determine the status of surface water quality in Florida, three categories of data – chemistry data, biological data, and fish consumption advisories – were evaluated to determine potential impairments. The level of impairment is defined in the Identification of Impaired Waters Rule (IWR), Section 62-303 of the Florida Administrative Code (FAC). The IWR is FDEP's methodology for determining whether waters should be included on the state's planning list and verified list. Potential impairments are determined by assessing whether a waterbody meets the criteria for inclusion on the planning list. Once a waterbody is on the planning list, additional data and information will be collected and examined to determine if the water should be included on the verified list.

The TMDL addressed in this document is being established pursuant to commitments made by the United States Environmental Protection Agency (USEPA) in the 1998 Consent Decree in the Florida TMDL lawsuit (Florida Wildlife Federation, et al. v. Carol Browner, et al., Civil Action No. 4: 98CV356-WS, 1998). That Consent Decree established a schedule for TMDL development for waters listed on Florida's USEPA approved 1998 section 303(d) list. The 2006 section 303(d) list identified numerous WBIDs in the Springs Coast Basin as not meeting water quality standards. After assessing all readily available water quality data, USEPA is responsible for developing a TMDL for the WBID depicted in Figure 2.1.

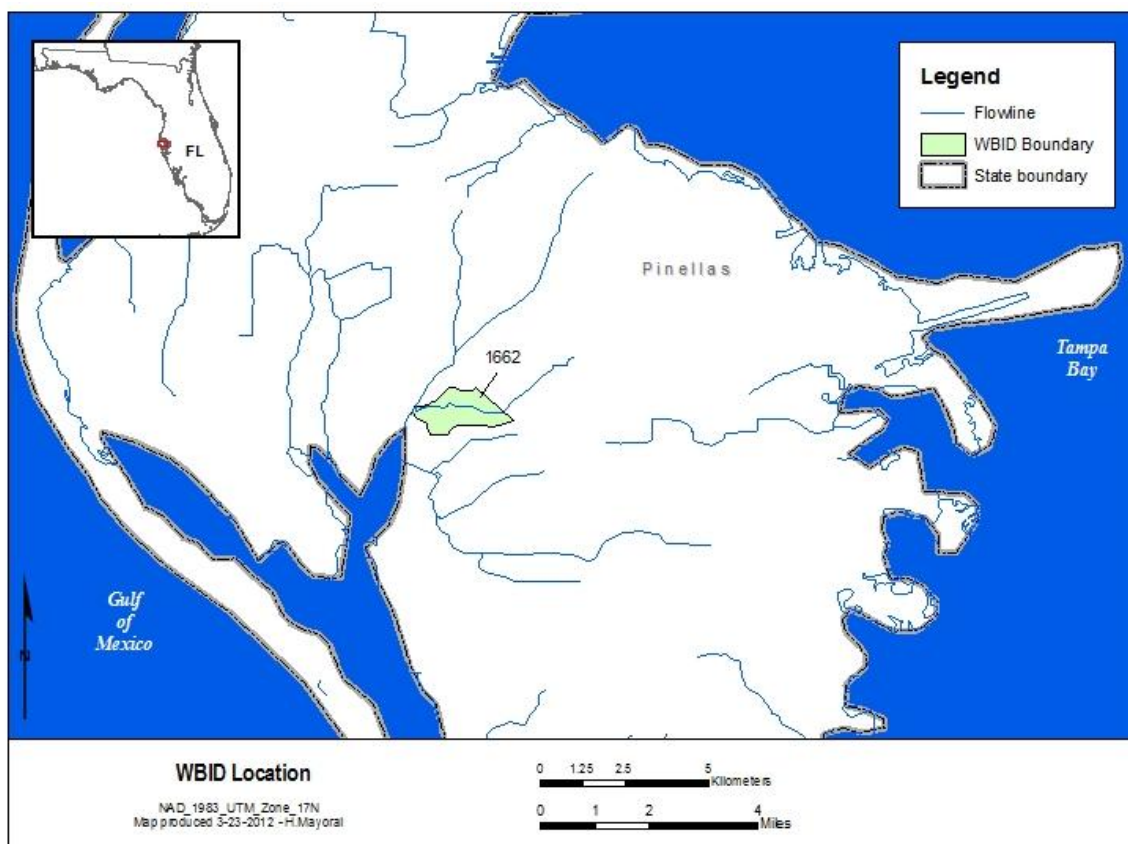


Figure 2.1 Location of impaired Pinellas Park Ditch No. 1, WBID 1662.

3.0 WATERSHED DESCRIPTION

The Springs Coast Basin is located along the west coast, beginning just south of the Withlacoochee River in Citrus County and extends to Gulfport, Florida in Pinellas County, although it does not include Tampa Bay. Within the watershed lies six major rivers, Crystal River, Homosassa River, Chassahowitza River, Weeki Wachee, the Anclote River, and the Pithlachascotee River; along with numerous springs and lakes (FDEP). The Brooksville Ridge marks the eastern boundary of the Springs Coast Basin, created by sands historically deposited during higher sea-levels, and which define the karst geology that is characteristic of the area (FDEP 2008).

Three physiographic regions with varying geology and topography are located within the Springs Coast Basin, the Coastal Swamps, the Gulf Coastal Lowlands, and the Brooksville Ridge. The Pinellas Park Ditches are located in the southern portion of the Springs Coast Basin in the Coastal Swamps region and the Gulf Coastal Lowlands (SWFWMD 2001). The Coastal Swamp region is characterized by poorly drained shallow soils that overlay the limestone of the Floridan aquifer. The Gulf Coast Lowlands are characterized by flat river valleys and rolling hills formed by aeolian deposited sands. Much of the Gulf Coastal Lowlands, including regions along the U.S. Highway 19 corridor, have been and continue to be intensively developed, although large sections of federally owned tracts of wetlands and swamps have been preserved (FDEP). The southwest region of the Springs Coast has fewer springs than the northern region, and the springs have relatively low flow volumes (FDEP 2008).

3.1 Climate

The Springs Coast Basin is located in Central Florida and experiences a humid subtropical climate with distinct wet (May to October) and dry (November to April) seasons, high rates of evapotranspiration, and climatic extremes of floods, droughts, and hurricanes. Seasonal rainfall patterns resemble the wet and dry season patterns of humid tropics. Central Florida receives an average of 46 inches of rain every year, of which 75% falls during the wet season, which coincides with hurricane season (USACE and SFWMD 2010). Average temperatures during the wet season are in the low-80s (°F) and in the dry season are in the mid-60s (°F) (NOAA).

3.2 Hydrologic characteristics

Pinellas Park Ditch No. 1 is located in the City of Pinellas Park in the western portion of Pinellas County. The area is heavily urbanized, and the watershed consists primarily of canals located in the developed areas. Pinellas Park Ditch No. 1 flows south west towards and drains to Cross Bayou, which empties into the Gulf of Mexico.

3.3 Land Use

The Pinellas Park Ditch No 1. WBID is located in a highly urbanized area. Developed land use accounts for 82 percent of the total land use within the WBID, of which 59 percent is high-intensity developed. Open water accounts for the next largest percentage of land use at 12 percent, while pasture and mixed forest comprise an additional 2 percent each. The remaining land use is made up of small regions of forest and nonforested wetlands, which are located at the base of the waterbody where it joins with the main ditch and exits into Cross Bayou.

The actual drainage area of the Pinellas Park Ditch varies from the WBID boundaries (Figure 3.2). The United States Geological Survey National Hydrography Dataset was used to delineate the drainage area. The land use composition for the drainage area was highly developed, and the amount

of developed land totaled 2,732 acres. Small percentages of land were in open water, forested, and forested wetland land uses.

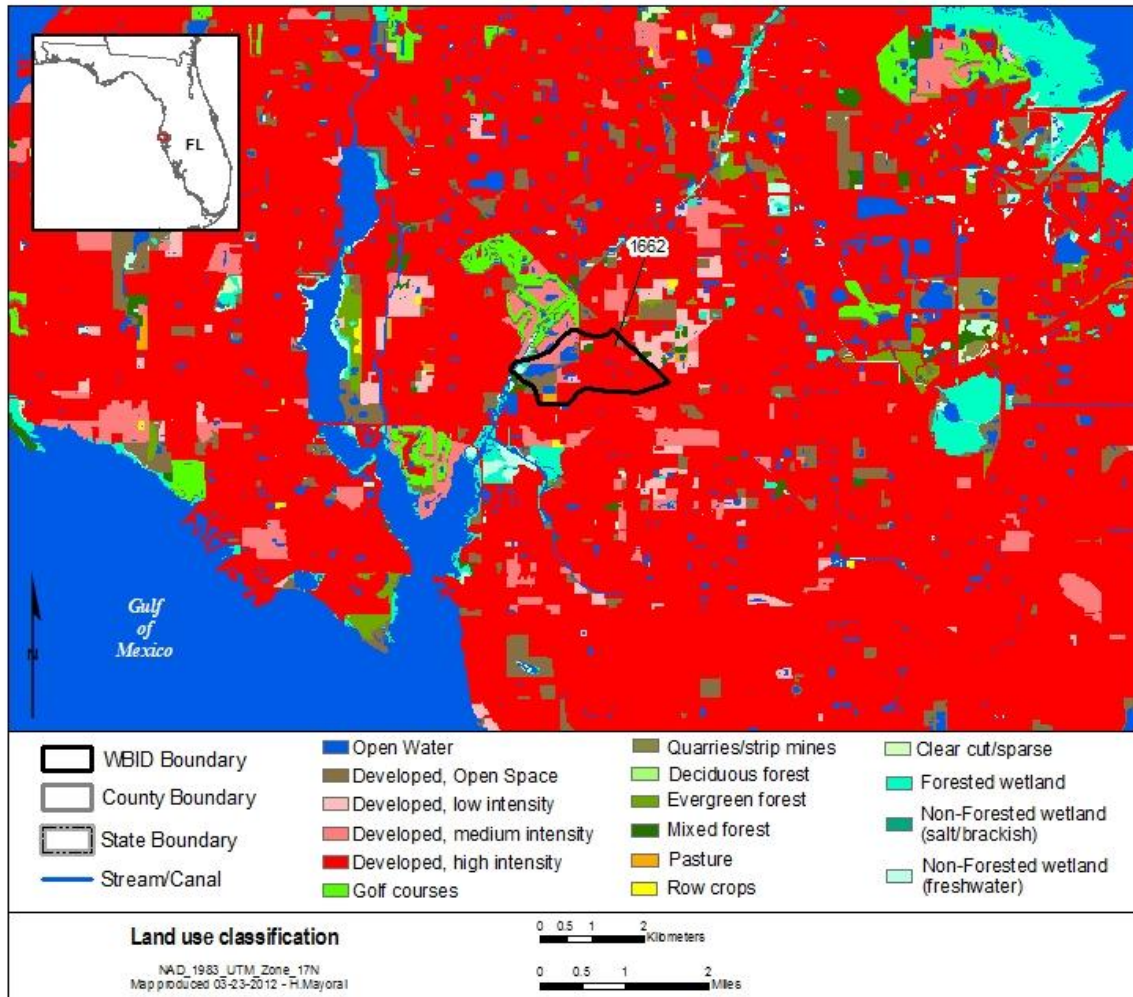


Figure 3.1 Land use in Pinellas Park Ditch No. 1.

Table 3.1 Land use distribution for the impaired WBID in the Pinellas Park Ditch basin.

Land Use Classification	1662	
	Acres	%
Evergreen Forest	0	0
Deciduous Forest	0	0
Mixed Forest	12	2
Forested Wetland	5	1
Non-Forested Wetland (Freshwater)	5	1
Open Water	65	12
Pasture	11	2
Row Crop	0	0
Golf Courses	0	0
Clear cut Sparse	0	0
Quarries Strip mines	0	0
Developed, Open Space	47	9
Developed, Low intensity	11	2
Developed, Medium intensity	64	12
Developed, High intensity	321	59
Totals	542	100

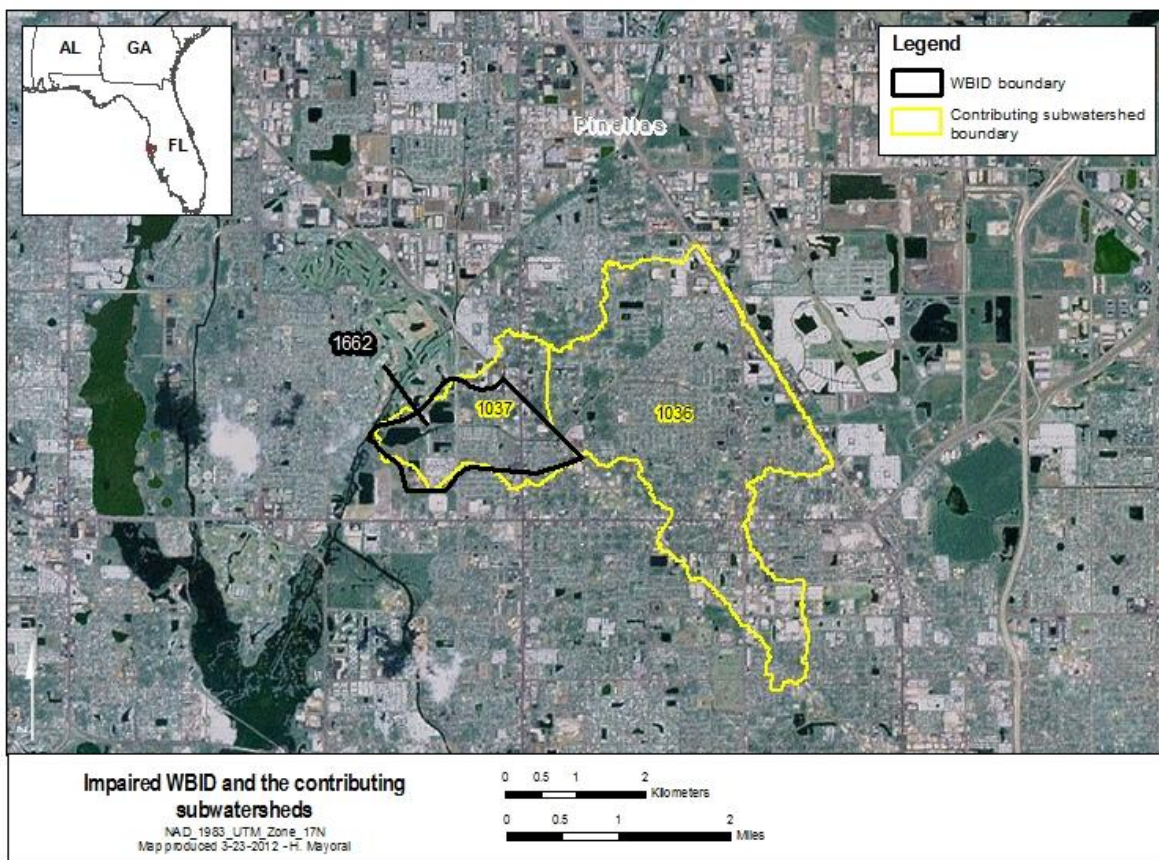


Figure 3.2 Aerial photo illustrating contributing subwatershed boundaries and impaired WBID boundary.

Table 3.2 Land use distribution for the contributing subwatersheds in the Pinellas Park Ditch No. 1 basin.

Land Use Classification	Subwatershed 1036		Subwatershed 1037		Total	
	Acres	%	Acres	%	Acres	%
Evergreen Forest	4	0%	4	1%	8	0%
Deciduous Forest	0	0%	0	0%	0	0%
Mixed Forest	23	1%	21	3%	44	1%
Forested Wetland	31	1%	6	1%	37	1%
Non-Forested Wetland (Freshwater)	3	0%	4	1%	7	0%
Open Water	58	3%	62	9%	120	4%
Pasture	6	0%	3	0%	10	0%
Row Crop	0	0%	0	0%	0	0%
Golf Courses	0	0%	1	0%	1	0%
Clear cut Sparse	0	0%	0	0%	0	0%
Quarries Strip mines	0	0%	0	0%	0	0%
Developed, Open Space	108	5%	48	7%	156	5%
Developed, Low intensity	169	7%	26	4%	196	7%
Developed, Medium intensity	28	1%	81	12%	110	4%
Developed, High intensity	1,842	81%	428	62%	2,270	77%
Totals	2,273	100%	685	100%	2,958	100%

4.0 WATER QUALITY STANDARDS/TMDL TARGETS

The TMDL reduction scenarios were done to achieve Florida's dissolved oxygen concentration of 5 mg/L and ensure balanced flora and fauna within these WBIDs or establish the TMDL to be consistent with a natural condition if the dissolved oxygen standard cannot be achieved.

4.1 Designated Uses

Florida has classified its waters based on the designated uses those waters are expected to support. Waters classified as Class I waters are designated for Potable Water Supply; Class II waters are designated for Shellfish Propagation or Harvesting, and Class III waters are designated for Recreation, Propagation and Maintenance of a Healthy, Well-Balanced Population of Fish and Wildlife. Designated use classifications are described in Florida's water quality standards at section 62-302.400, F.A.C.

Pinellas Park Ditch No. 1, WBID 1662, is classified as a Class III Marine water.

4.2 Water Quality Criteria

Water quality criteria for protection of all classes of waters are established in Section 62-302.530, F.A.C. Individual criteria should be considered in conjunction with other provisions in water quality standards, including Section 62-302.500 F.A.C., which established minimum criteria that apply to all waters unless alternative criteria are specified. Section 62-302.530, F.A.C. Several of the WBIDs addressed in this report were listed due to elevated concentrations of chlorophyll *a*. While FDEP does not have a streams water quality standard specifically for chlorophyll *a*, elevated levels of chlorophyll *a* are frequently associated with nonattainment of the narrative nutrient standard, which is described below.

4.3 Nutrient Criteria

In 1979, FDEP adopted a narrative criterion for nutrients. FDEP recently adopted numeric nutrient criteria for many Class III waters in the state, including streams, lakes, springs, and estuaries, which numerically interprets part of the state narrative criterion for nutrients. On November 30, 2012, EPA approved those criteria as consistent with the requirements of the CWA. Estuary specific criteria for a number of estuaries, as set out in 62-302.532(1), are effective for state law purposes. The remainder of the state criteria, however, are not yet effective for state law purposes.

In December 2010, EPA promulgated numeric nutrient criteria for Class I/III inland waters in Florida, including lakes and streams. On February 18, 2012, the federally promulgated criteria for lakes and springs were upheld by the U.S. District Court for the Northern District of Florida. Those criteria became effective on January 7, 2013. The Court invalidated the streams criteria and remanded those criteria back to EPA. EPA repropose the streams criteria on November 30, 2012.

Therefore, for lakes and springs in Florida, the applicable nutrient water quality criteria for CWA purposes are the federally promulgated criteria. For those estuaries identified in 62-302.532(1), the applicable nutrient water quality criteria for CWA purposes are FDEP's estuary criteria. For streams and the remaining estuaries in Florida, the applicable nutrient water quality standard for CWA purposes remains Florida's narrative nutrient criterion.

4.3.1 Narrative Nutrient Criteria

Florida's narrative nutrient criteria for Class I, II, and III waters provide:

The discharge of nutrients shall continue to be limited as needed to prevent violations of other standards contained in this chapter. Man induced nutrient enrichment (total nitrogen and total phosphorus) shall be considered degradation in relation to the provisions of Sections 62-302.300, 62-302.700, and 62-4.242. Section 62-302.530(47)(a), F.A.C.

In no case shall nutrient concentrations of a body of water be altered so as to cause an imbalance in natural populations of aquatic flora or fauna. Section 62-302.530(47)(b), F.A.C.

Chlorophyll and DO levels are often used to indicate whether nutrients are present in excessive amounts. The target for this TMDL is based on levels of nutrients necessary to prevent violations of Florida's DO criterion, set out below.

4.3.2 Inland Nutrient Criteria for streams

Florida's recently adopted numeric nutrient criteria interprets the narrative water quality criterion for nutrients in paragraph 62-302.530(48)(b), F.A.C. See section 62-302.531(2). While not yet effective as water quality criteria, the FDEP's numeric nutrient criteria represent the state's most recent interpretation of the second part of Florida's narrative criteria, set out at paragraph 62-302.530(47)(b), F.A.C. See section 62-302.531(2). Unless otherwise stated, where the EPA refers to the state nutrient rule in this TMDL, that rule is referenced as the state's interpretation of the narrative criterion. In addition, the first part of the narrative criteria, at paragraph 62-302.530(47)(a), F.A.C., also remains applicable to all Class I, II and III waters in Florida.

Florida's rule applies to streams. For streams that do not have a site specific criteria, Florida's rule provides for biological information to be considered together with nutrient thresholds to determine whether a waterbody is attaining 62-302.531(2)(c), F.A.C. The rule provides that the nutrient criteria are attained in a stream segment where information on chlorophyll a levels, algal mats or blooms, nuisance macrophyte growth, and changes in algal species composition indicates there are no imbalances in flora and either the average score of at least two temporally independent SCIs performed at representative locations and times is 40 or higher, with neither of the two most recent SCI scores less than 35, or the nutrient thresholds set forth in Table 1 below are achieved. See section 62-302.531(2)(c).

Florida's rule provides that numeric nutrient criteria are expressed as a geometric mean, and concentrations are not to be exceeded more than once in any three calendar year period. Section 62-302.200 (25)(e), F.A.C.

Table 4.1 Inland numeric nutrient criteria

Nutrient Watershed Region	Total Phosphorus Nutrient Threshold	Total Nitrogen Nutrient Threshold
Panhandle West	0.06 mg/L	0.67 mg/L
Panhandle East	0.18 mg/L	1.03 mg/L

North Central	0.30 mg/L	1.87 mg/L
Peninsular	0.12 mg/L	1.54 mg/L
West Central	0.49 mg/L	1.65 mg/L
South Florida	No numeric nutrient threshold. The narrative criterion in paragraph 62-302.530(47)(b), F.A.C., applies.	No numeric nutrient threshold. The narrative criterion in paragraph 62-302.530(47)(b), F.A.C., applies.

4.3.3 Inland Nutrient Criteria for estuaries with effective criteria

Numeric criteria for estuaries are expressed as either concentration-based estuary interpretations that are open water, area-wide averages or as load per million cubic meters of freshwater inflow that are the total load of that nutrient to the estuary divided by the total volume of freshwater inflow to that estuary. The criteria are set out at 62-302.532(1).

4.3.4 Inland Nutrient Criteria for lakes

Federal water quality criteria for lakes set out at 40 CFR 131.43(c)(1). The criteria are expressed as concentrations of chlorophyll *a*, total phosphorus, and total nitrogen as follows:

Lake Color and Alkalinity	Chl- <i>a</i> (mg/L)*	TN (mg/L)	TP (mg/L)
Colored Lakes (Long-term Color > 40 Platinum Cobalt Units (PCU))	0.020	1.27 [1.27-2.23]	0.05 [0.05-0.16]
Clear Lakes, High Alkalinity (Long-term Color ≤ 40 PCU and Alkalinity > 20 mg/L CaCO ₃)	0.020	1.05 [1.05-1.91]	0.03 [0.03-0.09]
Clear Lakes, Low Alkalinity (Long-term Color ≤ 40 PCU and Alkalinity ≤ 20 mg/L CaCO ₃)	0.006	0.51 [0.51-0.93]	0.01 [0.01-0.03]

* For a given waterbody, the annual geometric mean of chlorophyll *a*, TN or TP concentrations shall not exceed the applicable criterion concentration more than once in a three-year period.

4.3.5 Springs Nutrient Criteria

The numeric criteria for spring is 0.35 mg/L of nitrate-nitrite as an annual geometric mean, not to be exceeded more than once in any three year period.

4.4 Dissolved Oxygen Criteria

Numeric criteria for DO are expressed in terms of minimum and daily average concentrations. While FDEP has adopted revised DO criteria for freshwaters, these revisions have not yet been submitted to EPA for review. Therefore, the applicable criterion for Clean Water Act purposes remains subsection 62-302.530(30), F.A.C.

For Class I and Class III freshwaters, subsection 62-302.530(30) provides as follows:

Shall not be less than 5.0 mg/L. Normal daily and seasonal fluctuations above these levels shall be maintained. [FAC 62-302.530 (30)]

For Class III marine waters, subsection 62-302.530(30) provides as follows:

Shall not average less than 5.0 mg/L in a 24-hour period and shall never be less than 4.0 mg/L. Normal daily and seasonal fluctuations above these levels shall be maintained. [FAC 62-302.530 (30)]

4.5 Biochemical Oxygen Demand Criteria

Biochemical Oxygen Demand (BOD) shall not be increased to exceed values which would cause dissolved oxygen to be depressed below the limit established for each class and, in no case, shall it be great enough to produce nuisance conditions. [FAC 62-302.530 (11)]

4.6 Natural Conditions

In addition to the standards for nutrients, DO, and BOD described above, Florida's standards include provisions that address waterbodies which do not meet the standards due to natural background conditions.

Florida's water quality standards provide a definition of natural background:

"Natural Background" shall mean the condition of waters in the absence of man-induced alterations based on the best scientific information available to the Department. The establishment of natural background for an altered waterbody may be based upon a similar unaltered waterbody or on historical pre-alteration data. [FAC 62-302.200(19)]

Florida's water quality standards also provide that:

Pollution which causes or contributes to new violations of water quality standards or to continuation of existing violations is harmful to the waters of this State and shall not be allowed. Waters having water quality below the criteria established for them shall be protected and enhanced. However, the Department shall not strive to abate natural conditions. [FAC 62-302.300(15)]

5.0 WATER QUALITY ASSESSMENT

The WBID addressed in this report was listed as not attaining its designated use on Florida's 1998 303(d) list for dissolved oxygen and nutrients. To determine impairment, an assessment of available data was conducted. The source for current ambient monitoring data was the Impaired Waters Rule

(IWR) data Run 44, using data ranging January 1, 2002 to December 31, 2010. The IWR database contains data from various sources within the state of Florida, including the WMDs and counties.

5.1 Water Quality Data

Analysis of water quality data for the WBID is documented on Table 5.1, and Figure 5.1 shows the locations of the water quality monitoring stations within WBID 1662. Water quality data for the WBID can be found below in Figures 5.2 through 5.6.

5.1.1 Dissolved Oxygen

There are several factors that affect the concentration of DO in a waterbody. Oxygen can be introduced by wind, diffusion, photosynthesis, and additions of higher DO water (e.g. from tributaries). DO concentrations are lowered by processes that use up oxygen from the water, such as respiration and decomposition, and by additions of water with lower DO (e.g. swamp or groundwater). Natural DO levels are a function of water temperature, water depth and velocity, and relative contributions of groundwater. Decomposition of organic matter, such as dead plants and animals, also consume DO. The minimum measured DO concentration in WBID 1662 was 2.36 mg/L, and the maximum measured concentration was 8.61 mg/L. The mean DO concentration was 4.86 mg/L.

5.1.2 Biological Oxygen Demand

BOD is a measure of the amount of oxygen used by bacteria as they stabilize organic matter. The process can be accelerated when there is an overabundance of nutrients, increasing the aerobic bacterial activity in a waterbody. In turn, the levels of DO can become depleted, eliminating oxygen essential for biotic survival, and potentially causing fish kills. Additionally, BOD is used as an indicator to determine the presence and magnitude of organic pollution from sources such as septic tank leakage, fertilizer runoff, and wastewater effluent. The mean BOD measurement in WBID 1662 was 1.78 mg/L. The maximum concentration was 3.10 mg/L and the minimum was 0.77 mg/L.

5.1.3 Nutrients

Excessive nutrients in a waterbody can lead to overgrowth of algae and other aquatic plants such as phytoplankton, periphyton and macrophytes. This process can deplete oxygen in the water, adversely affecting aquatic life and potentially restricting recreational uses such as fishing and boating. For the nutrient assessment the monitoring data for total nitrogen, total phosphorus and chlorophyll a are presented. The current standards for nutrients are narrative criteria. The purpose of the nutrient assessment is to present the range, variability and average conditions for the WBID.

5.1.3.1 Total Nitrogen

Total Nitrogen (TN) is comprised of nitrate (NO₃), nitrite (NO₂), organic nitrogen and ammonia nitrogen (NH₄). Though nitrogen is a necessary nutrient required for the growth of most plants and animals, not all forms are readily used or metabolized. Increased levels of organic nitrogen can occur from the decomposition of aquatic life or from sewage, while inorganic forms are generally from erosion and fertilizers. Nitrates are components of industrial fertilizers, yet can also be naturally present in soil, and are converted to nitrite by microorganisms in the environment. Surface runoff from agricultural lands can increase the natural presence of nitrates in the environment and can lead to eutrophication. In WBID 1662, the measured total nitrogen minimum concentration was 0.50 mg/L, and the maximum measured concentration was 1.53 mg/L. The total nitrogen mean was 1.23 mg/L.

5.1.3.2 Total Phosphorus

In natural waters, total phosphorus exists in either soluble or particulate forms. Dissolved phosphorus includes inorganic and organic forms, while particulate phosphorus is made up of living and dead plankton, and adsorbed, amorphous, and precipitated forms. Inorganic forms of phosphorus include orthophosphate and polyphosphates, though polyphosphates are unstable and convert to orthophosphate over time. Orthophosphate is both stable and reactive, making it the form most used by plants. Excessive phosphorus can lead to overgrowth of algae and aquatic plants, the decomposition of which uses up oxygen from the water. The measured total phosphorus minimum concentration in WBID 1662 was 0.08 mg/L, and the maximum concentration was 0.31 mg/L. The mean total phosphorus concentration was 0.16 mg/L.

5.1.3.3 Chlorophyll-a

Chlorophyll is the green pigment in plants that allows them to create energy from light. In a water sample, chlorophyll is indicative of the presence of algae, and chlorophyll-*a* is a measure of the active portion of total chlorophyll. Corrected chlorophyll refers to chlorophyll-*a* measurements that are corrected for the presence of pheophytin, a natural degradation product of chlorophyll that can interfere with analysis because it has an absorption peak in the same spectral region. It is used as a proxy indicator of water quality because of its predictable response to nutrient availability. Increases in nutrients can potentially lead to blooms in phytoplankton biomass, affecting water quality and ecosystem health. The corrected chlorophyll-*a* maximum measurement in WBID 1662 was 37.0 µg/L and the minimum was 1.0 µg/L. The mean chlorophyll-*a* concentration was 5.81 µg/L. The minimum measurements of 1.0 µg/L all occurred during 2006 and were at the detection limit. All other chlorophyll *a* taken during other years were above the detection limit, and several of the measurements were greater than 20 µg/L.

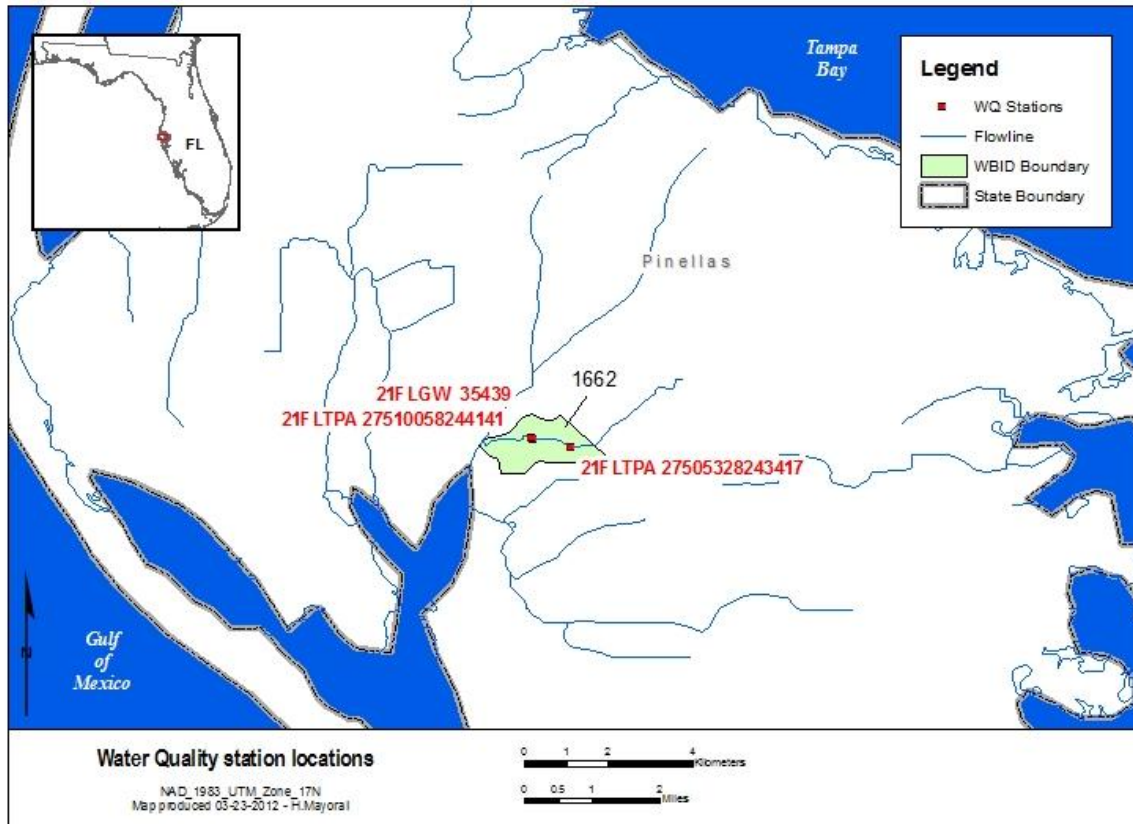


Figure 5.1 Water quality monitoring station locations for WBID 1662.

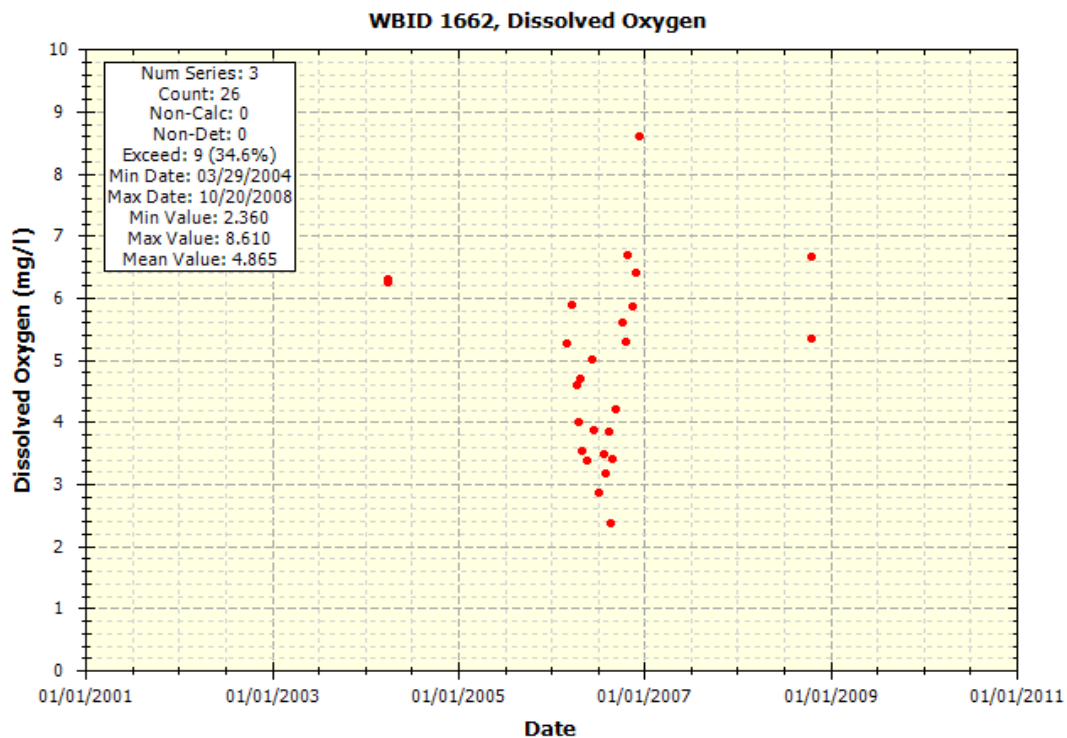


Figure 5.2 Dissolved oxygen concentrations for WBID 1662.

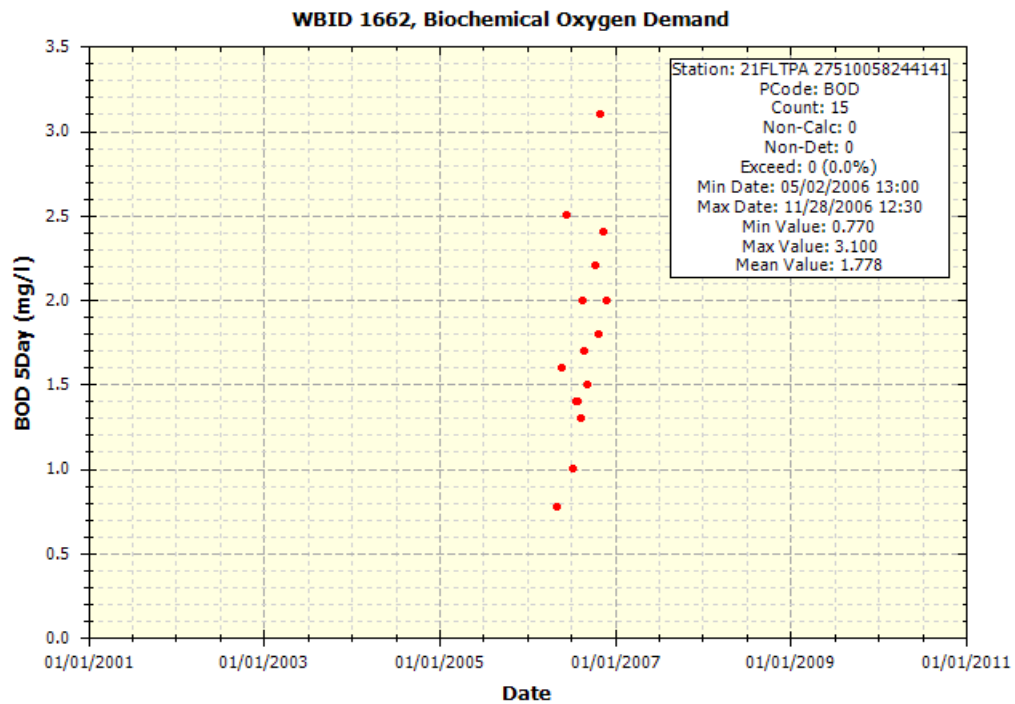


Figure 5.3 Biochemical oxygen demand concentrations for WBID 1662.

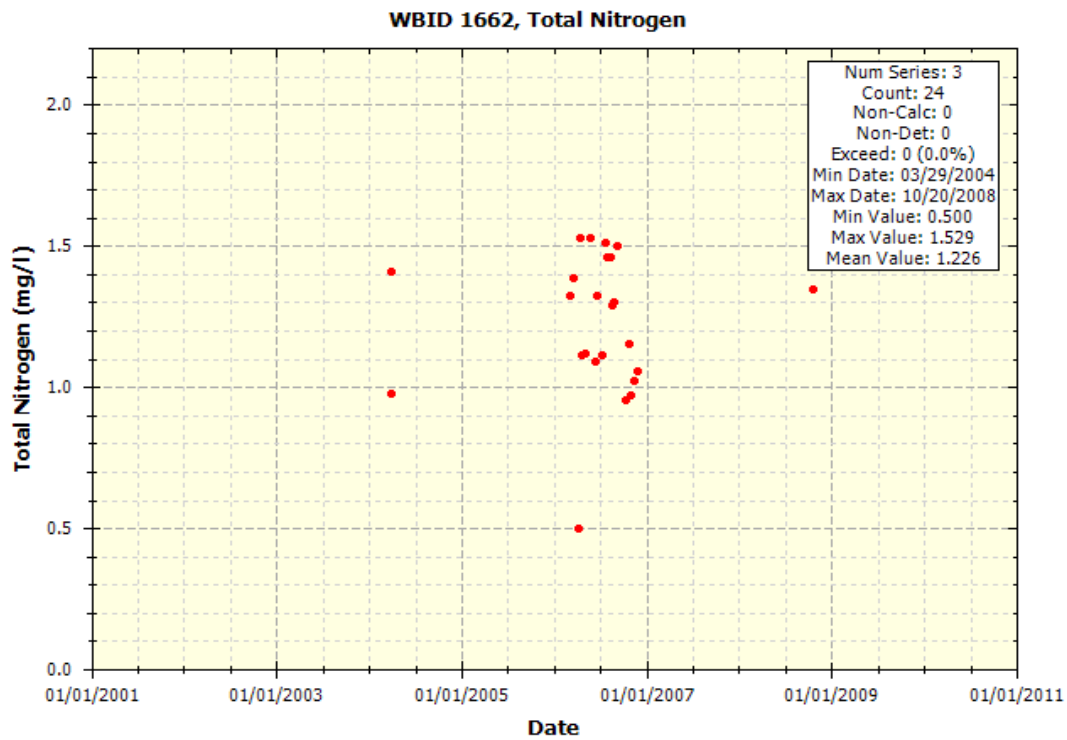


Figure 5.4 Total nitrogen concentrations for WBID 1662.

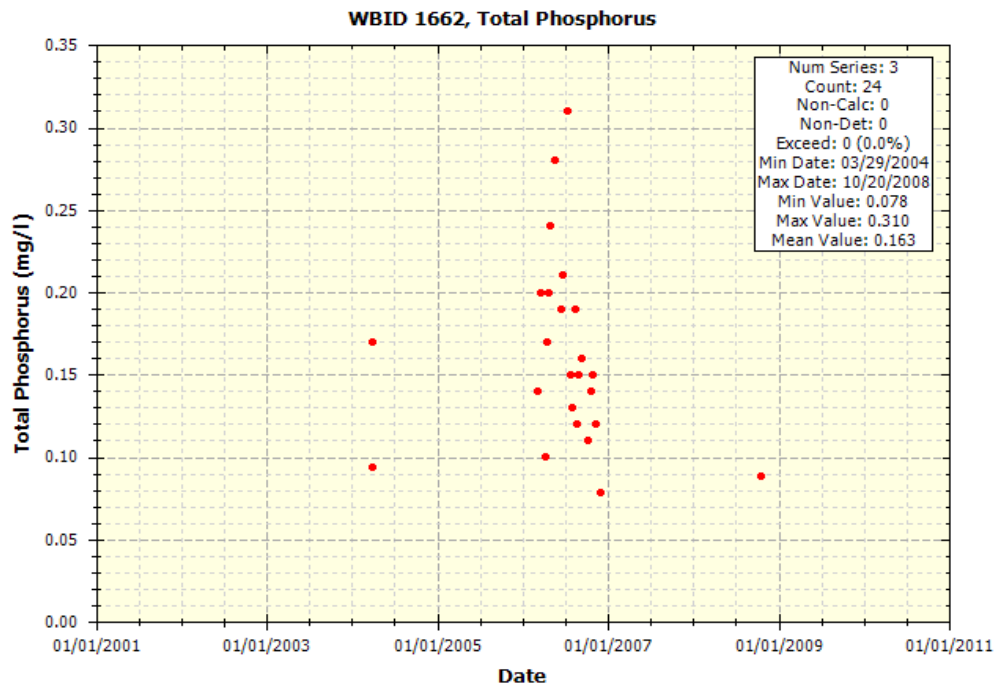


Figure 5.5 Total phosphorus concentrations for WBID 1662.

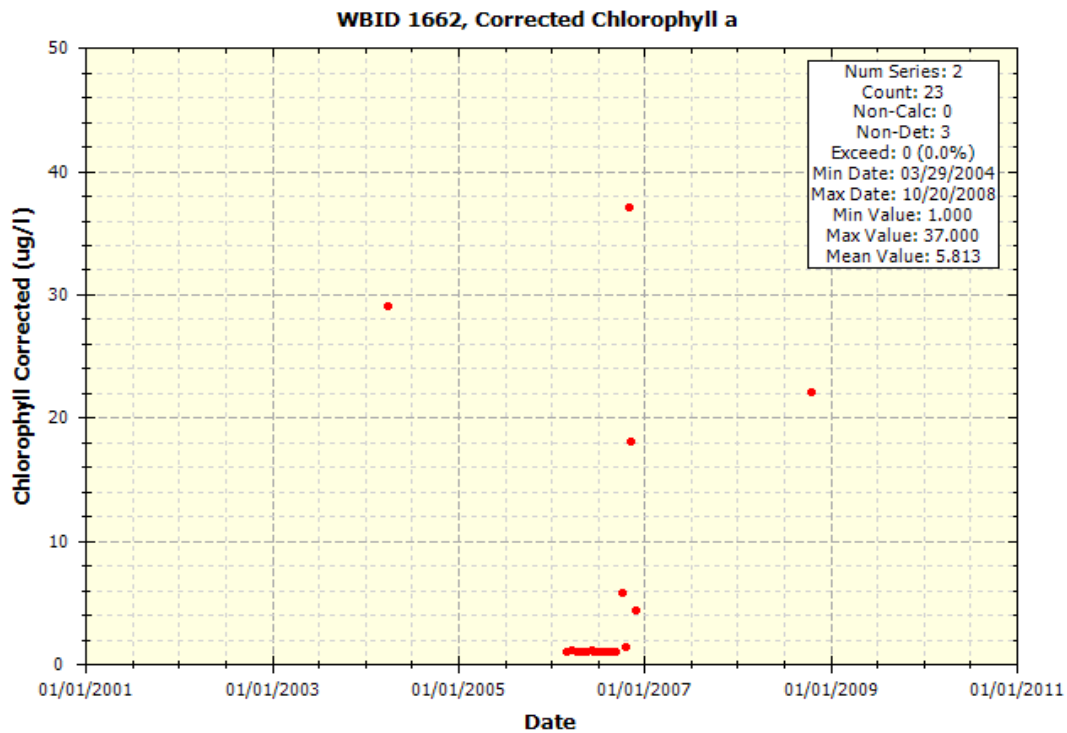


Figure 5.6 Corrected chlorophyll a concentrations for WBID 1662.

Table 5.1 Water quality data for impaired WBID in the Park Pinellas Ditch No. 1 basin.

Parameter	Stats	WBID
		1662
BOD, 5 Day, 20°C (mg/L)	# of obs	15
	min	0.77
	max	3.10
	mean	1.78
	Geomean	1.68
DO, Analysis by Probe (mg/L)	# of obs	26
	min	2.36
	max	8.61
	mean	4.86
	Geomean	4.65
Nitrogen, Total (mg/L as N)	# of obs	24
	min	0.50
	max	1.53
	mean	1.23
	Geomean	1.20
Phosphorus, Total (mg/L as P)	# of obs	24
	min	0.08
	max	0.31
	mean	0.16
	Geomean	0.15
Chlorophyll I-A- corrected (µg/L)	# of obs	23
	min	1.00

	max	37.00
	mean	5.81
	Geomean	2.07

6.0 SOURCE AND LOAD ASSESSMENT

An important part of the TMDL analysis is the identification of source categories, source subcategories, or individual sources of pollutants in the watershed and the amount of loading contributed by each of these sources. Sources are broadly classified as either point or nonpoint sources. Nutrients can enter surface waters from both point and nonpoint sources.

6.1 Point Sources

A point source is defined as a discernible, confined, and discrete conveyance from which pollutants are or may be discharged to surface waters. Point source discharges of industrial wastewater and treated sanitary wastewater must be authorized by National Pollutant Discharge Elimination System (NPDES) permits. NPDES permitted discharges include continuous discharges such as wastewater treatment facilities as well as some stormwater driven sources such as municipal separate storm sewer systems (MS4s), certain industrial facilities, and construction sites over one acre.

6.1.1 Wastewater/Industrial Permitted Facilities

A TMDL wasteload allocation (WLA) is given to wastewater and industrial NPDES permitted facilities discharging to surface waters within an impaired watershed. There are no NPDES-permitted facilities discharging to WBID 1662.

6.1.2 Stormwater Permitted Facilities/MS4s

MS4s are point sources also regulated by the NPDES program. According to 40 CFR 122.26(b)(8), an MS4 is “a conveyance or system of conveyances (including roads with drainage systems, municipal streets, catch basins, curbs, gutters, ditches, man-made channels, or storm drains):

(i) Owned or operated by a State, city, town, borough, county, parish, district, association, or other public body (created by or pursuant to State law)...including special districts under State law such as a sewer district, flood control district or drainage district, or similar entity, or an Indian tribe or an authorized Indian tribal organization, or a designated and approved management agency under section 208 of the Clean Water Act that discharges into waters of the United States;

(ii) Designed or used for collecting or conveying storm water;

(iii) Which is not a combined sewer; and

(iv) Which is not part of a Publicly Owned Treatment Works.”

MS4s may discharge nutrients and other pollutants to waterbodies in response to storm events. In 1990, USEPA developed rules establishing Phase I of the NPDES stormwater program, designed to prevent harmful pollutants from being washed by stormwater runoff into MS4s (or from being dumped directly into the MS4) and then discharged from the MS4 into local waterbodies. Phase I of the program required operators of “medium” and “large” MS4s (those generally serving populations of 100,000 or greater) to implement a stormwater management program as a means to control

polluted discharges from MS4s. Approved stormwater management programs for medium and large MS4s are required to address a variety of water quality related issues including roadway runoff management, municipal owned operations, hazardous waste treatment, etc.

Phase II of the rule extends coverage of the NPDES stormwater program to certain “small” MS4s. Small MS4s are defined as any MS4 that is not a medium or large MS4 covered by Phase I of the NPDES stormwater program. Only a select subset of small MS4s, referred to as “regulated small MS4s”, requires an NPDES stormwater permit. Regulated small MS4s are defined as all small MS4s located in “urbanized areas” as defined by the Bureau of the Census, and those small MS4s located outside of “urbanized areas” that are designated by NPDES permitting authorities.

In October 2000, USEPA authorized FDEP to implement the NPDES stormwater program in all areas of Florida except Indian tribal lands. FDEP’s authority to administer the NPDES program is set forth in Section 403.0885, Florida Statutes (FS). The three major components of NPDES stormwater regulations are:

- MS4 permits that are issued to entities that own and operate master stormwater systems, primarily local governments. Permittees are required to implement comprehensive stormwater management programs designed to reduce the discharge of pollutants from the MS4 to the maximum extent practicable.
- Stormwater associated with industrial activities, which is regulated primarily by a multisector general permit that covers various types of industrial facilities. Regulated industrial facilities must obtain NPDES stormwater permit coverage and implement appropriate pollution prevention techniques to reduce contamination of stormwater.
- Construction activity general permits for projects that ultimately disturb one or more acres of land and which require the implementation of stormwater pollution prevention plans to provide for erosion and sediment control during construction.

Stormwater discharges conveyed through the storm sewer system covered by the permit are subject to the WLA of the TMDL. Any newly designated MS4s will also be required to achieve the percent reduction allocation presented in this TMDL. The MS4 permits for WBID 1662 are listed on Table 6.2.

Table 6.2 MS4 Permits by WBID.

WBID	Segment Name	Phase	Facility Number	Permittee	Co-Permittee(s)
1662	Pinellas Park Ditch No 1 (Tidal Segment)	I C	FLS000005	Pinellas County	City of Pinellas Park Florida Department of Transportation

6.2 Nonpoint Sources

Nonpoint sources of pollution are diffuse sources that cannot be identified as entering a waterbody through a discrete conveyance at a single location. For nutrients, these sources include runoff of agricultural fields, golf courses, and lawns, septic tanks, and residential developments outside of MS4 areas. Nonpoint source pollution generally involves a buildup of pollutants on the land surface that wash off during rain events and as such, represent contributions from diffuse sources, rather than

from a defined outlet. Potential nonpoint sources are commonly identified, and their loads estimated, based on land cover data. Most methods calculate nonpoint source loadings as the product of the water quality concentration and runoff water volume associated with certain land use practices. The mean concentration of pollutants in the runoff from a storm event is known as the event mean concentration. Figure 3.1 provides a map of the land use in the WBID. Table 3.1 lists the land use distribution for the WBID.

The following sections are organized by land use. Each section provides a description of the land use, the typical sources of nutrient loading (if applicable), and typical total nitrogen and total phosphorus event mean concentrations.

6.2.1 Urban Areas

Urban areas include land uses such as residential, industrial, extractive and commercial. Land uses in this category typically have somewhat high total nitrogen event mean concentrations and average total phosphorus event mean concentrations. Nutrient loading from MS4 and non-MS4 urban areas is attributable to multiple sources including stormwater runoff, leaks and overflows from sanitary sewer systems, illicit discharges of sanitary waste, runoff from improper disposal of waste materials, leaking septic systems, and domestic animals.

In 1982, Florida became the first state in the country to implement statewide regulations to address the issue of nonpoint source pollution by requiring new development and redevelopment to treat stormwater before it is discharged. The Stormwater Rule, as outlined in Chapter 403 FS, was established as a technology-based program that relies upon the implementation of Best Management Practices (BMPs) that are designed to achieve a specific level of treatment (i.e., performance standards) as set forth in Chapter 62-40, FAC.

Florida's stormwater program is unique in having a performance standard for older stormwater systems that were built before the implementation of the Stormwater Rule in 1982. This rule states: "the pollutant loading from older stormwater management systems shall be reduced as needed to restore or maintain the beneficial uses of water." [FAC 62-40-.432(2)(c)]

Nonstructural and structural BMPs are an integral part of the State's stormwater programs. Nonstructural BMPs, often referred to as "source controls", are those that can be used to prevent the generation of nonpoint source pollutants or to limit their transport off-site. Typical nonstructural BMPs include public education, land use management, preservation of wetlands and floodplains, and minimization of impervious surfaces. Technology-based structural BMPs are used to mitigate the increased stormwater peak discharge rate, volume, and pollutant loadings that accompany urbanization.

Urban, residential, and commercial developments are a significant nonpoint source of nutrients and oxygen-demanding substances in the Pinellas Park Ditch No. 1 watershed. In the contributing subwatersheds, approximately 93 percent of the land use, or 2,732 acres, is developed. Of the developed land uses, 2,270 acres are categorized as high intensity development.

Onsite Sewage Treatment and Disposal Systems (Septic Tanks)

As stated above, leaking septic tanks or onsite sewage treatment and disposal systems (OSTDs) can contribute to nutrient loading in urban areas. Water from OSTDs is typically released to the ground through on-site, subsurface drain fields or boreholes that allow the water from the tank to percolate (usually into the surficial aquifers) and either transpire to the atmosphere through surface vegetation

or add to the flow of shallow ground water. When properly sited, designed, constructed, maintained, and operated, OSTDs are a safe means of disposing of domestic waste. The effluent from a well-functioning OSTD receives natural biological treatment in the soil and is comparable to secondarily treated wastewater from a sewage treatment plant. When not functioning properly, OSTDs can be a source of nutrients, pathogens, and other pollutants to both ground water and surface water.

The State of Florida Department of Health publishes data on new septic tank installations and the number of septic tank repair permits issued for each county in Florida. Table 6.3 summarizes the cumulative number of septic systems installed in Pinellas County since the 1970 census and the total number of repair permits issued for the ten years between 1999-2000 and 2009-2010 (FDOH 2009). The data do not reflect septic tanks removed from service. Leaking septic systems could be a relevant source of organic and nutrient loading in the watershed.

Table 6.3 County estimates of Septic Tanks and Repair Permits.

County	Number of Septic Tanks (1970-2008)	Number of Repair Permits Issued (2000-2010)
Pinellas	23,869	3,015

Note: Source: <http://www.doh.state.fl.us/environment/ostds/statistics/ostdsstatistics.htm>

6.2.2 Pastures

Pastures include cropland and improved and unimproved pasturelands, such as non-tilled grasses woodland pastures, feeding operations, nurseries and vineyards; as well as specialty farms. Agricultural activities, including runoff of fertilizers or animal wastes from pasture and cropland and direct animal access to streams, can generate nutrient loading to streams. The highest total nitrogen and total phosphorus event mean concentrations are associated with agricultural land uses.

The USDA National Agricultural Statistics Service (NASS) compiles Census of Agriculture data by county for virtually every facet of U.S. agriculture (USDA NASS 2007). According to the 2007 Census of Agriculture, there were 56 farms within Pinellas County (Table 6.4). The amount of acreage in those farms being fertilized with commercial fertilizer, lime and soil conditioners was as little as 152 acres over 46 farms, with the remaining ten farms fertilizing with manure totaling 48 acres. Livestock counts of cattle and pigs in Pinellas County are provided in Table 6.5. Due to agricultural census data being collected at the county level, the extent to which these values pertain to agricultural fields within the impaired watershed is not specified.

Only 10 acres in the Pinellas Park Ditch No. 1 contributing watershed are in pasture land uses, and no acreage is in row crop land uses. For this reason, non-point source pollution from agricultural land use likely has a negligible negative impact on water quality in the watershed.

Table 6.4 Agricultural Census Data for Commercially and manure fertilized farms in Pinellas County, Florida.

County	Commercial		Manure	
	Number of Farms	Number of Acres	Number of Farms	Number of Acres
Pinellas	46	152	10	48

Table 6.5 Agricultural Census Data for Livestock in Pinellas County, Florida.

County	Livestock	Number of Farms	Number of Animals
Pinellas	Cattles and Calves	14	130
	Hogs and Pigs	2	-

Note: 1. A farm is defined as any place from which \$1,000 or more of agricultural products were produced and sold, or normally would have been sold, during the census year.

6.2.3 Clear cut/Sparse

The clear cut/sparse land use classification includes recent clear cuts, areas of sparse vegetation or herbaceous dry prairie, shrub and brushland, other early successional areas, and mixed rangeland. Event mean concentrations for clear cut/sparse are about average for total nitrogen and low for total phosphorus. There is no clearcut/sparse land use within the contributing subwatersheds to WBID 1662.

6.2.4 Forests

Upland forests include flatwoods, oak, various types of hardwoods, conifers and tree plantations. Wildlife, located within forested areas, deposit their feces onto land surfaces where it can be transported to nearby streams during storm events. Generally, the pollutant load from wildlife is assumed to represent background concentrations. Event mean concentrations for upland forests are low for both total nitrogen and total phosphorus. Combined forest land use comprises 2 percent of the land use distribution within the contributing subwatersheds.

6.2.5 Water and Wetlands

Water and Wetlands have very low event mean concentrations down to zero and account for the largest land use class after developed land uses. Approximately 164 acres are classified as wetland or open water land uses in the contributing subwatersheds, comprising 5 percent of the total contributing land uses.

6.2.6 Quarries/Strip mines

Land use classification includes quarries, strip mines, exposed rock and soil, fill areas, reclaimed lands, and holding ponds. Event mean concentrations for some barren lands tend to be higher in total nitrogen. There are no quarries/strip mines within the contributing subwatersheds.

7.0 ANALYTICAL APPROACH

In the development of a TMDL there needs to be a method for relating current loadings to the observed water quality problem. This relationship could be: statistical (regression for a cause and effect relationship), empirical (based on observations not necessarily from the waterbody in question) or mechanistic (physically and/or stochastically based) that inherently relate cause and effect using physical and biological relationships.

Mechanistic models were used in the development of the Pinellas Park Ditch No. 1 TMDL to relate the physical and biological relationships. A dynamic watershed model was used to predict the quantity of water and pollutants associated with runoff from rain events. The watershed model was linked to a hydrodynamic model that simulated tidal influences in the river. Both models were linked to a water quality simulation model that integrated the loadings and flow from the watershed model with flow from the hydrodynamic model to predict the water quality in the receiving waterbodies.

The period of simulation that was considered in the development of this TMDL is January 1, 2002 to December 31, 2009. The models were used to predict time series for BOD, TN, TP, and DO. The models were calibrated to current conditions and were then used to predict improvements in water quality as function of reductions in loadings.

7.1 *Mechanistic Models*

7.1.1 Loading Simulation Program C++ (LSPC)

LSPC is the Loading Simulation Program in C++, a watershed modeling system that includes streamlined Hydrologic Simulation Program Fortran (HSPF) algorithms for simulating hydrology, sediment, and general water quality overland as well as a simplified stream fate and transport model. LSPC is derived from the Mining Data Analysis System (MDAS), which was originally developed by USEPA Region 3 (under contract with Tetra Tech) and has been widely used for TMDLs. In 2003, the USEPA Region 4 contracted with Tetra Tech to refine, streamline, and produce user documentation for the model for public distribution. LSPC was developed to serve as the primary watershed model for the USEPA TMDL Modeling Toolbox. LSPC was used to simulate runoff (flow, biological oxygen demand, total nitrogen, total phosphorus and dissolved oxygen) from the land surface using a daily timestep for current and natural conditions. LSPC provided tributary flows and temperature to the EFDC estuary models and tributary water quality concentrations to WASP7 estuary models.

An LSPC model was utilized to estimate the nutrient loads within and discharged from the Pinellas Park Ditch watershed. The LSPC model utilized data inputs from the larger Crystal Watershed model developed for the Florida Numeric Nutrient Criteria (USEPA 2012a and USEPA 2012b).

In order to evaluate the contributing sources to a waterbody and to represent the spatial variability of these sources within the watershed model, the contributing drainage area was represented by a series of sub-watersheds for each of the models. The sub-watersheds for the Crystal Watershed model were developed using the 12-digit hydrologic unit code (HUC12) watershed data layer and the Geological Survey (USGS) National Hydrograph Dataset (NHD). The Crystal Watershed model was then further redefined into smaller subwatersheds for the Pinellas Park Ditch model using the NHD catchments and USGS National Elevation Dataset Digital Elevation Model (DEM) (Figure 7.1).

The LSPC model has a representative reach defined for each sub-watershed, and the main channel stem within each sub-watershed was used as the representative reach. The characteristics for each reach include the length and slope of the reach, the channel geometry and the connectivity between the sub-watersheds. Length and slope data for each reach was obtained using the USGS DEM and NHD data.

The attributes supplied for each reach were used to develop a function table (FTABLE) that describes the hydrology of the stream reach by defining the functional relationship between water depth, surface area, water volume, and outflow in the segment. The assumption of a fixed depth, area, volume, outflow relationship rules out cases where the flow reverses direction or where one reach influences another upstream of it in a time-dependent way. LSPC does not model the tidal flow in the low-lying estuaries, and therefore the Crystal Watershed model was calibrated to non-tidally influenced USGS gages, and the hydrodynamic parameterization from the Crystal Watershed was used in the Pinellas Park Ditch model. The Pinellas Park Ditch Watershed model was linked to the EFDC and WASP models to simulate the areas of the estuary that were tidally influenced.

The watershed model uses land use data as the basis for representing hydrology and nonpoint source loadings. The FDEP Level III Florida Land Use, specifically the Southwest Florida Water Management District (SWFWMD) 2006 dataset, was used to determine the land use representation. The National Landuse Coverage Dataset (NLCD) was used to develop the impervious land use representations.

The SWFWMD coverage utilized a variety of land use classes which were grouped and re-classified into 18 land use categories: beaches/dune/mud, open water, utility swaths, developed open space, developed low intensity, developed medium intensity, developed high intensity, clear-cut/sparse, quarries/strip mines, deciduous forest, evergreen forest, mixed forest, golf courses, pasture, row crop, forested wetland, non-forested wetland (salt/brackish), and non-forested wetland (freshwater). The LSPC model requires division of land uses in each sub-watershed into separate pervious and impervious land units. The 2006 NLCD 2006 percent impervious coverage was used to determine the percent impervious area associated with each land use category. Any impervious areas associated with utility swaths, developed open space, and developed low intensity, were grouped together and placed into a new land use category named low intensity development impervious. Impervious areas associated with medium intensity development and high intensity development were kept separate and placed into two new categories for medium intensity development impervious and high intensity development impervious, respectively. Finally, any impervious area not already accounted for in the three developed impervious categories, were grouped together into a fourth new category for all remaining impervious land use.

Soil data was obtained from the Soil Survey Geographic Database (SSURGO). The database was produced and distributed by the Natural Resources Conservation Service (NRCS) - National Cartography and Geospatial Center (NCGC). The SSURGO data was used to determine the total area that each hydrologic soil group covered within each sub-watershed. The sub-watersheds were represented by the hydrologic soil group that had the highest percentage of coverage within the boundaries of the sub-watershed.

In the watershed model, nonpoint source loadings and hydrological conditions are dependent on weather conditions. Hourly data from weather stations within the boundaries of, or in close proximity to, the sub-watersheds were applied to the watershed model. A weather data forcing file was generated in ASCII format (*.air) for each meteorological station used in the hydrological

evaluations in LSPC. Each meteorological station file contained atmospheric data used in modeling the hydrological processes. These data included precipitation, air temperature, dew point temperature, wind speed, cloud cover, evaporation, and solar radiation. These data are used directly, or calculated from the observed data. The Pinellas Park Ditch Watershed model weather stations contained data through 2009.

No flow calibration stations were located in Pinellas Park Ditch or in the nearby vicinity. For this reason, the Pinellas Park Ditch model used the same parameters that were used in the larger Crystal Watershed (USEPA 2012b). The model parameters that were adjusted in the Crystal Watershed model consist of evapotranspiration, infiltration, upper and lower zone storage, groundwater storage, losses to the deep groundwater system, and Manning's n . The calibration of the LSPC watershed hydrology model involved comparing simulated streamflows to the USGS flow stations. The calibration of the hydrologic parameters was performed from January 1, 1997, through December 31, 2009.

The Pinellas Park Ditch watershed model also used the Crystal Watershed model water quality parameters. Water quality results were compared with measured data from IWR 44 in the Pinellas Park Ditch WBID. Water quality parameters, including nutrient accumulation and washoff, were adjusted for a better calibration. The water quality calibration for the Pinellas Park Ditch Model is shown in Figure 7.2 and Figure 7.6.

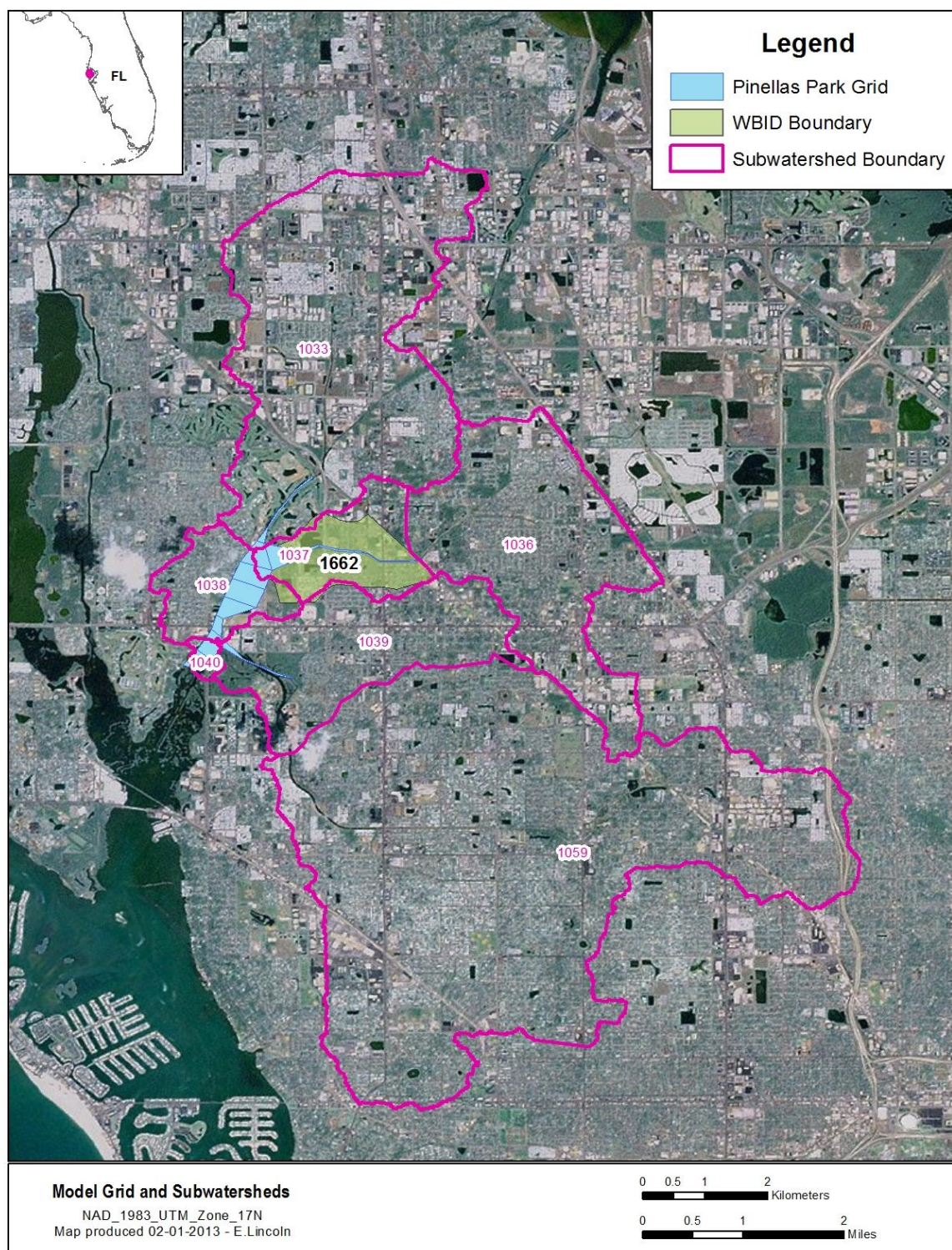


Figure 7.1 LSPC subwatershed boundaries and WASP model grid for Pinellas Park Ditch No. 1

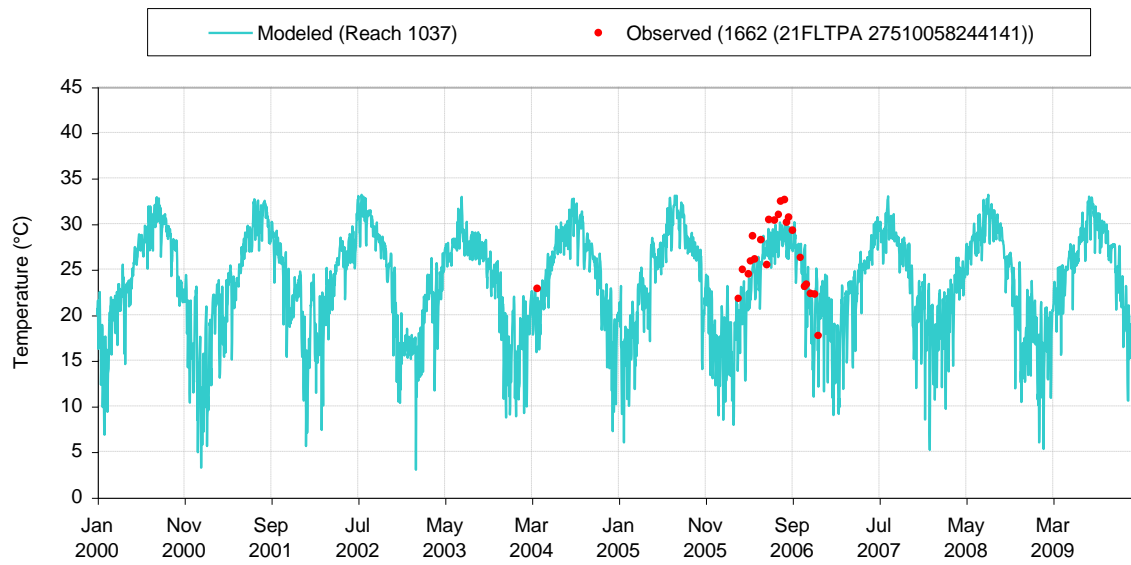


Figure 7.2 Modeled vs. Observed Temperature (°C) at 1662 (21FLTPA 27510058244141)

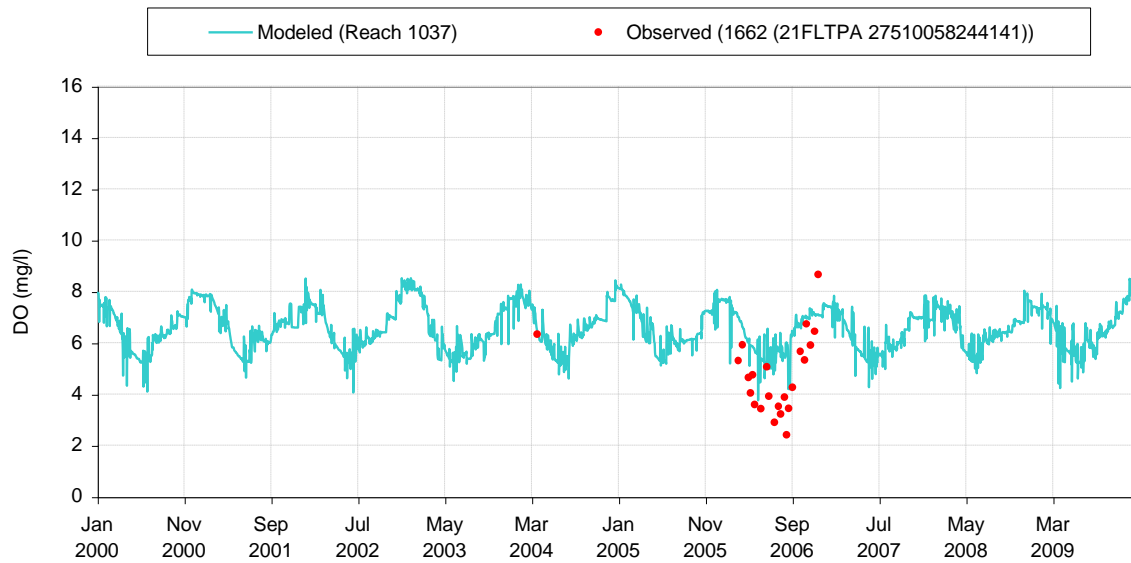


Figure 7.3 Modeled vs. Observed DO (mg/l) at 1662 (21FLTPA 27510058244141)

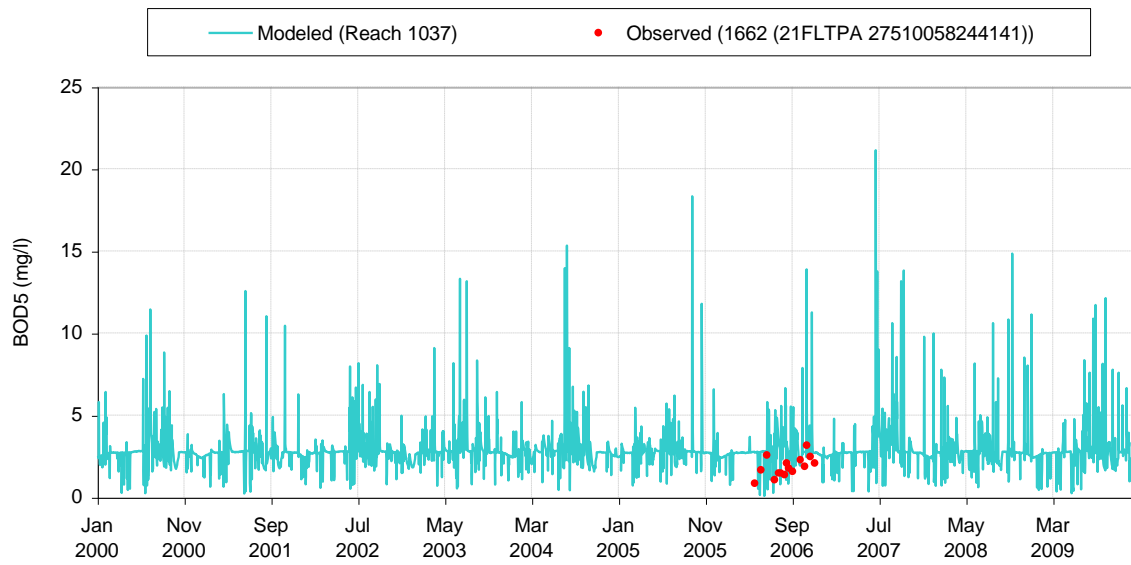


Figure 7.4 Modeled vs. Observed BOD5 (mg/l) at 1662 (21FLTPA 27510058244141)

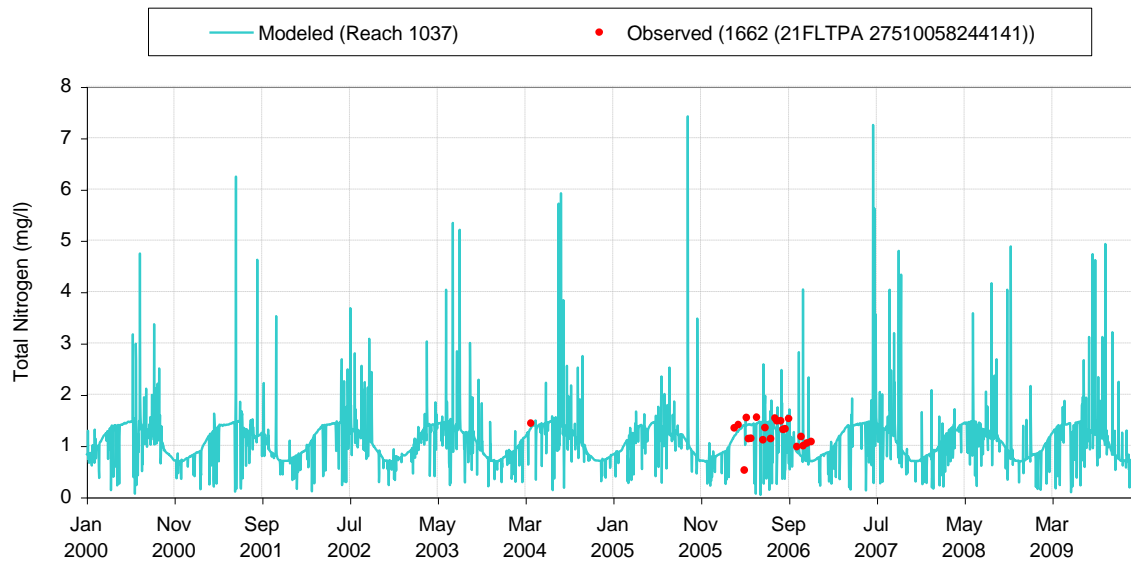


Figure 7.5 Modeled vs. Observed Total Nitrogen (mg/l) at 1662 (21FLTPA 27510058244141)

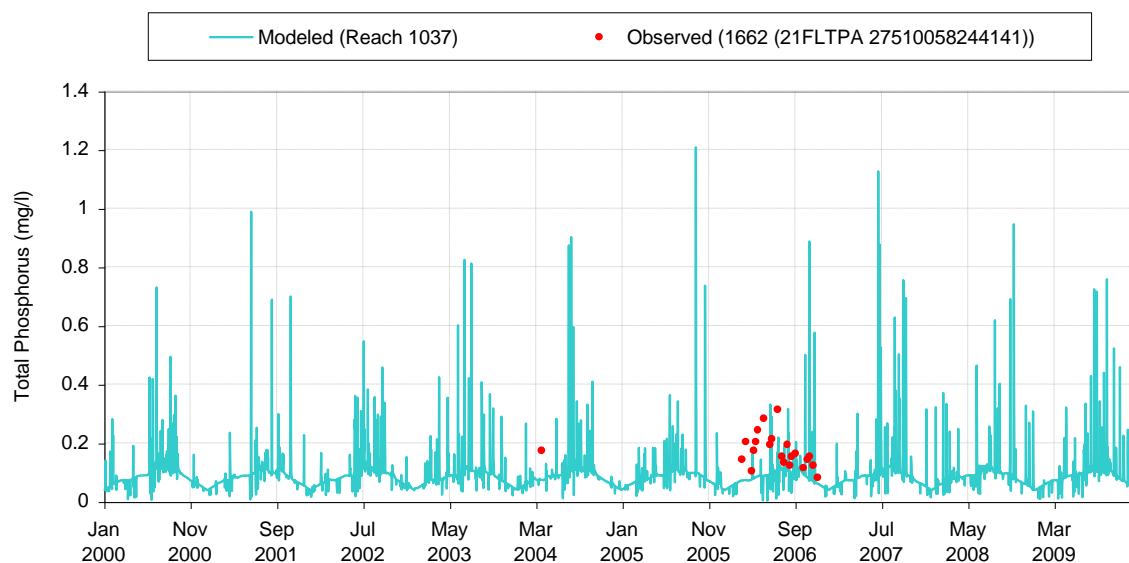


Figure 7.6 Modeled vs. Observed Total Phosphorus (mg/l) at 1662 (21FLTPA 27510058244141)

7.1.2 Environmental Fluids Dynamic Code (EFDC)

The EFDC model is a part of the USEPA TMDL Modeling Toolbox due to its application in many TMDL-type projects. As such, the code has been peer reviewed and tested and has been freely distributed and supported by Tetra Tech. EFDC was developed by Dr. John Hamrick (Hamrick 1992) and is currently supported by Tetra Tech for USEPA Office of Research and Development (ORD), USEPA Region 4, and USEPA Headquarters. The models, tools, and databases in the TMDL Modeling Toolbox are continually updated and upgraded through TMDL development in Region 4. EFDC is a multifunctional, surface-water modeling system, which includes hydrodynamic, sediment contaminant, and eutrophication components. The EFDC model is capable of 1, 2, and 3-dimensional spatial resolution. The model employs a curvilinear-orthogonal horizontal grid and a sigma or terrain following vertical grid.

The EFDC hydrodynamic model can run independently of a water quality model. The EFDC model simulates the hydrodynamic and constituent transport and then writes a hydrodynamic linkage file for a water quality model such as the Water Quality Analysis Program (WASP7) model. This model linkage, from EFDC hydrodynamics to WASP water quality, has been applied on many USEPA Region 4 projects in support of TMDLs and has been well tested (Wool et al. 2003).

The EFDC model was utilized to simulate three-dimensional circulation dynamics of hydrodynamic state variables (water surface elevation, salinity, and temperature) in the Pinellas Park Ditch model. An orthogonal, curvilinear grid system consisting of 3995 horizontal cells and 4 equally spaced vertical layers was developed for the Big Bend EFDC model. The grid was developed using Gulf of Mexico bathymetry data. The large grid was reduced in size and scale for the Pinellas Park Ditch EFDC model. Bathymetry was unavailable for the inland, tidally influenced streams, and channel slope from the USGS digital elevation model was used to estimate slope within the channel. To

conduct numerical simulation of the hydrodynamics and water quality, the Pinellas Park Ditch grid was generated using the GEFDC and VOGG programs, which were used to compute cell widths and lengths (Figure 7.7).

The EFDC model predicts water surface elevation, salinity, and temperature, in response to a set of multiple factors: wind speed and direction, freshwater discharge, tidal water level fluctuation, rainfall, surface heat flux, and temperature and salinity associated with boundary fluxes. Hourly measurements of atmospheric pressure, dry and wet bulb atmospheric temperatures, rainfall rate, wind speed and direction, and fractional cloud cover were obtained from data collected at two WBAN stations, Apalachicola and Clearwater, for 2002 through 2009. Solar short wave radiation was calculated using the CE-Qual-W2 method.

The Big Bend model used hourly water surface elevation time series data from the National Oceanic and Atmospheric Administration (NOAA) tidal stations to simulate tides at the open boundary. Observed temperature data at water quality stations were used to simulate the temperature at the open boundaries, and average salinity in the Gulf of Mexico was used to simulate salinity. The Big Bend Estuary was calibrated to measured NOAA tidal stations, and the Big Bend model was used to simulate the open boundary conditions in the Pinellas Park Ditch model. The upstream inland boundary grid cells received LSPC simulated watershed discharges and LSPC loadings. Freshwater enters the system via two paths, including direct reach inflows and overland runoff. LSPC model results were converted to EFDC model input files QSER.INP and TSER.INP.

Because there were no NOAA tidal stations located within Pinellas Park Ditch, water surface elevation within the modeled cells could not be directly calibrated. Salinity measurements from IWR44 data were used to review the Pinellas Park Ditch estuary EFDC calibration. Following model review, the salinity and temperature parameters were adjusted accordingly (Figure 7.8 and Figure 7.9).



Figure 7.7 LSPC subwatershed boundaries, estuary model grid, and calibration station for Pinellas Park Ditch No. 1

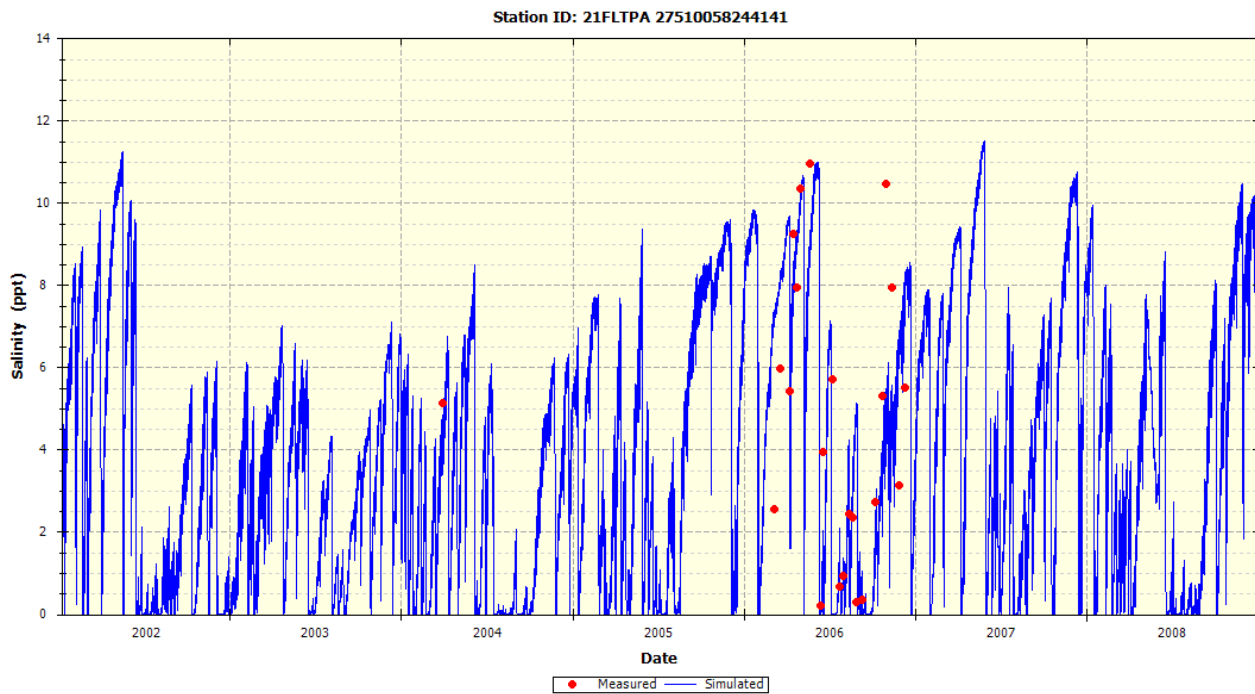


Figure 7.8 Simulated salinity verse measured salinity in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

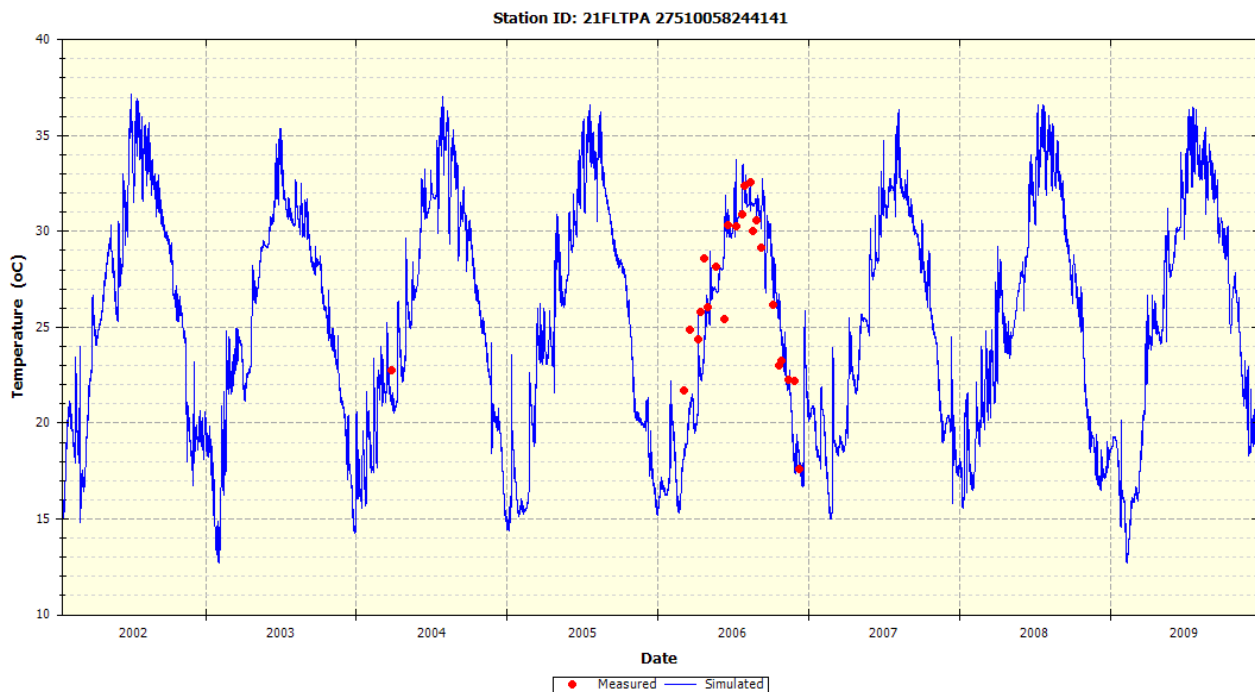


Figure 7.9 Simulated temperature (°) verse measured temperature in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

7.1.3 Water Quality Analysis Simulation Program (WASP7)

The Water Quality Analysis Simulation Program Version 7.4.1 (WASP7) is an enhanced Windows version of the USEPA Water Quality Analysis Simulation Program (WASP) (Di Toro et al., 1983; Connolly and Winfield, 1984; Ambrose, R.B. et al., 1988), with upgrades to the user's interface and the model's capabilities. The major upgrades to WASP have been the addition of multiple BOD components, addition of sediment diagenesis routines, and addition of periphyton routines. The hydrodynamic file generated by EFDC is compatible with WASP7 and it transfers segment volumes, velocities, temperature and salinity, as well as flows between segments. The time step is set in WASP7 based on the hydrodynamic simulation.

WASP7 helps users interpret and predict water quality responses to natural phenomena and man-made pollution for various pollution management decisions. WASP7 is a dynamic compartment-modeling program for aquatic systems, including both the water column and the underlying benthos. The time-varying processes of advection, dispersion, point and diffuse mass loading and boundary exchange are represented in the basic program. Water quality processes are represented in special kinetic subroutines that are either chosen from a library or written by the user. WASP is structured to permit easy substitution of kinetic subroutines into the overall package to form problem-specific models. WASP7 comes with two such models, TOXI for toxicants and EUTRO for conventional water quality.

Because the LSPC model simulated TN, TP, and BOD and the WASP model simulated TN and its speciation, TP and its speciation, and CBOD, the water quality concentrations from LSPC were adjusted for WASP simulation prior to being input into the WASP model. TN was speciated into nitrate-nitrite (NOX), ammonia (NH₄), and organic nitrogen (ON), and TP was speciated into orthophosphate (PO₄) and organic phosphorus (OP). Water quality data in the Cross Bayou Canal was reviewed to determine the ratio of NOX, NH₄, and ON in TN, and the ratio of PO₄ and OP in TP. The in-stream BOD loads from LSPC were converted to ultimate CBOD using an f-ratio of 1.5.

WASP7 utilized the same EFDC grid cell for the Pinellas Park Ditch model. Water quality loading from the LSPC model was used to simulate loads coming from rivers and streams into the estuary. Initially, default values of the rates and constants provided by WASP7 were used. Following initial set-up, calibration parameters from the Tampa Bay estuary model were used to populate rates and constants. The WASP7 model was run iteratively with updated rates and constants until model agree well with data in terms of temporal trends and magnitudes. The Pinellas Park Ditch model was calibrated at a station located within the impaired WBID, and the results are shown in Figure 7.10 through Figure 7.14.

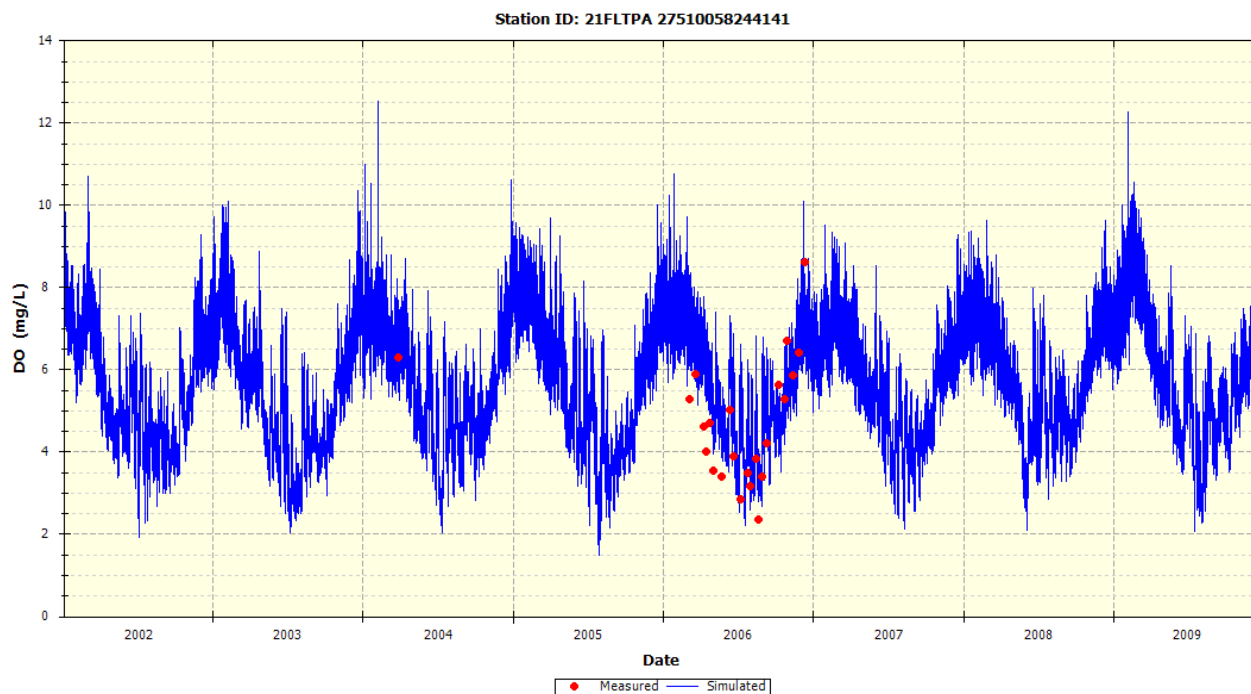


Figure 7.10 Simulated dissolved oxygen (mg/L) verse measured dissolved oxygen in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

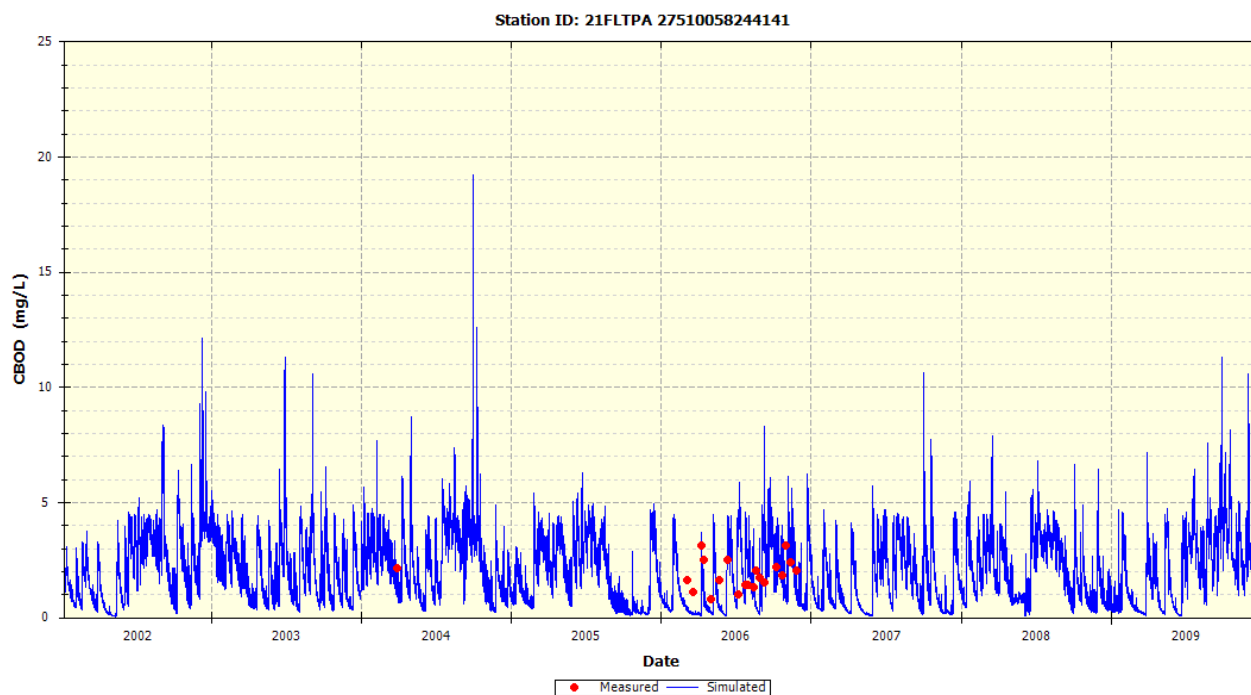


Figure 7.11 Simulated cBOD (mg/L) verse measured cBOD in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

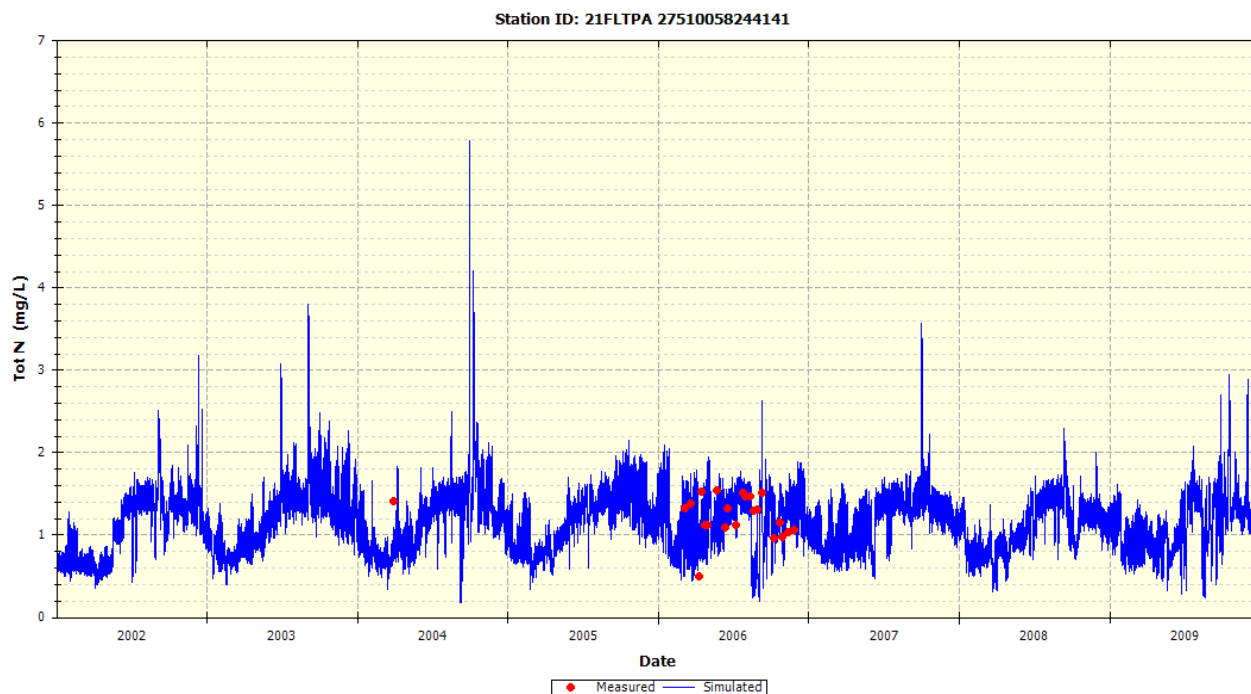


Figure 7.12 Simulated total nitrogen (mg/L) verse measured total nitrogen in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

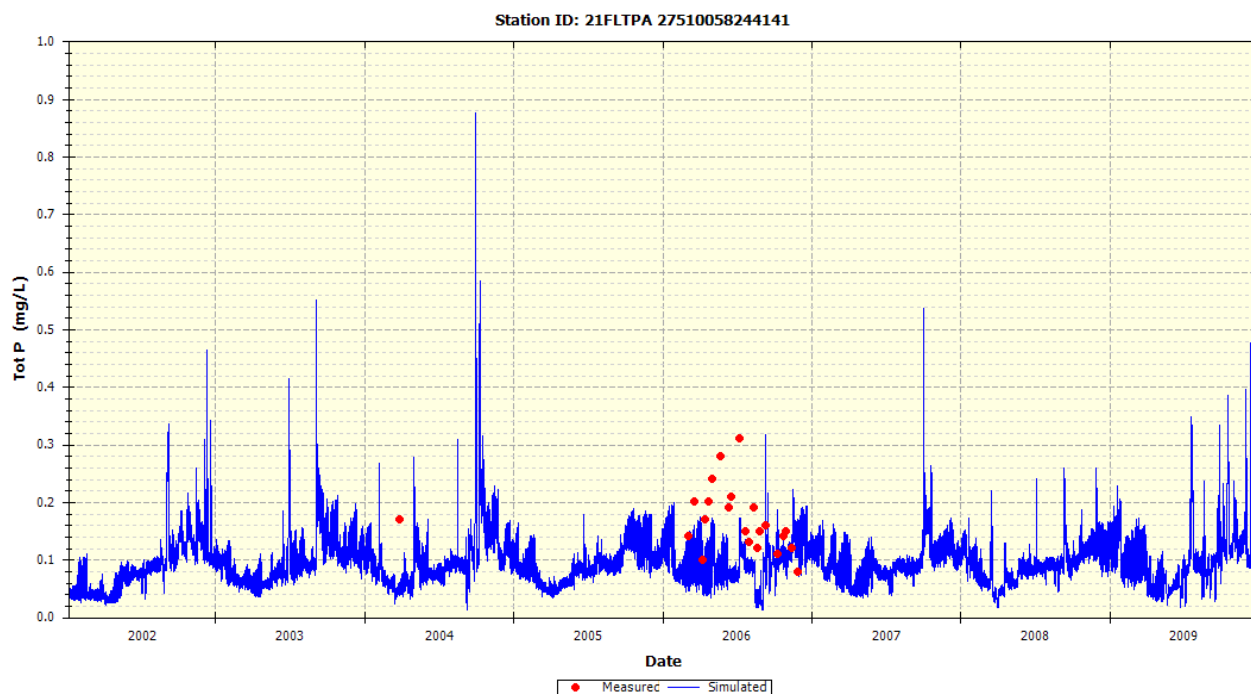


Figure 7.13 Simulated total phosphorus (mg/L) verse measured total phosphorus in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

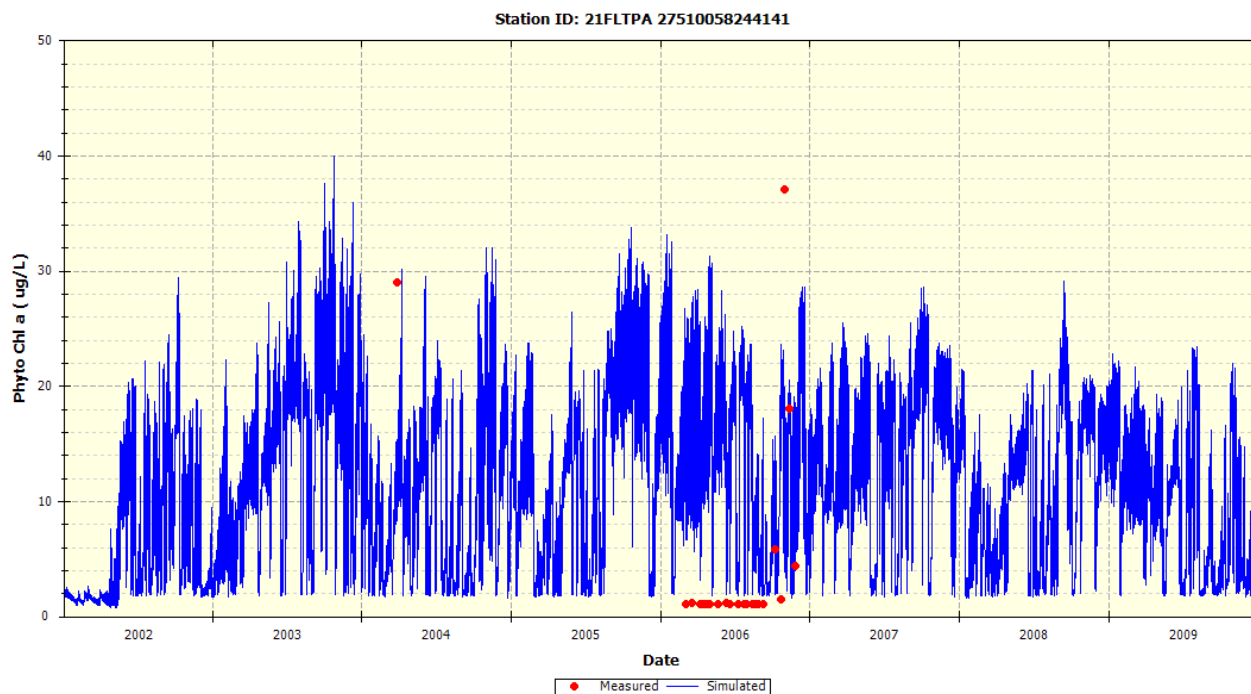


Figure 7.14 Simulated chlorophyll a (ug/L) verse measured chlorophyll a in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

7.1.4 Pinellas Park Ditch Basin Mechanistic Model

To model the Pinellas Park Ditch No. 1 basin, the relatively large subwatersheds in the Crystal Watershed Model were re-delineated using the USGS NHD catchments. The subwatersheds used in the development of the Pinellas Park Ditch model were re-delineated to a smaller size, and the same parameterization as the larger Crystal model was used. Additionally, the original EFDC model grid was re-drawn at a smaller scale and extended up into the Pinellas Park Ditch No. 1 basin (Figure 7.1). The Tampa Estuary model parameterization was used in the Pinellas Park Ditch model, and outputs from the adjacent Tampa Estuary model grid cells were used as the open boundary conditions in the Pinellas Park Ditch model. The Pinellas Park Ditch LSPC model outputs were input into the inland cells.

A WASP7 grid was developed for the Pinellas Park Ditch model and parameterized using the Tampa Estuary model WASP7 parameters. The hydrodynamic simulation from the Pinellas Park Ditch model was input into the WASP7 model. Loadings from the LSPC model were input into the inland cells, and EFDC water quality outputs were used for the WASP7 open boundary cells. When re-delineated to the small scale, it was determined that loadings from the original Crystal model were low in the Pinellas Park Ditch basin. Loadings from the watershed were increased for the Pinellas Park Ditch modeling effort for successful calibration.

7.2 Scenarios

Two modeling scenarios were developed and evaluated in this TMDL determination: a current condition and a natural condition scenario. Concentrations and loadings were evaluated to determine

if DO concentrations in the natural condition scenario could meet the DO standard, and the impact of nutrients on the DO concentrations. The results from the scenarios were used to develop the TMDL.

7.2.1 Current Condition

The current condition scenario evaluated current hydrologic and water quality conditions in the watershed, specifically water quality concentration and loadings at the outlet of 1662. Modeled calibration results, which are daily average concentrations, of the model is presented in Figure 7.2 through Figure 7.6. The current condition annual average concentrations for the Pinellas Park Ditch No. 1 WBID are presented in Table 7.1. The current condition simulation was used to determine the base loadings for the WBID. These base loadings (Table 7.2), when compared with the TMDL scenarios, were used to determine the percent reduction in nutrient loads that will be needed to achieve water quality standards.

Table 7.1 Current condition concentrations in the impaired WBID in Pinellas Park Ditch basin.

Parameter	WBID 1662
Total nitrogen (mg/L)	1.16
Total phosphorus (mg/L)	0.10
CBOD (mg/L)	1.99
DO (mg/L)	5.70

Table 7.2 Current condition loadings in the impaired WBID in Pinellas Park Ditch basin.

Parameter	WBID 1409	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	5,905
Total phosphorus (mg/L)	--	697
BOD5 (mg/L)	--	16,620

7.2.2 Natural Condition

The natural condition scenario was developed to estimate water quality conditions if there was no impact from anthropogenic sources. Landuses that were associated with anthropogenic activities (urban, agriculture, transportation, barren lands and rangeland) were converted to upland forests or forested wetlands based on the current ratio of forest and wetland landuses in the model. Additionally, following the initial natural condition scenario run, sediment oxygen demand (SOD) was revised by using the following formula: $SOD_{\text{revised}} = (\text{Avg Chla}_{\text{natural}} / \text{Avg Chla}_{\text{existing}}) * SOD$. The lower, revised SOD represents the change expected in SOD following excessive nutrient

removal from the system. The natural condition water quality predictions are presented in Table 7.3 and 7.4.

The purpose of the natural conditions scenario was to determine whether water quality standards could be achieved without abating the naturally occurring loads from the watershed. The natural condition modeling scenario indicated that the daily average DO standard of 5 mg/L is not achievable under natural conditions, indicating that low DO is a naturally occurring phenomenon in the WBID 1662. Figure 7.15 through Figure 7.19 provide the natural condition scenario modeled parameters for WBID 1662. Figure 7.20 provides the cumulative distribution function of DO concentrations for both the modeled existing condition and natural condition results. The cumulative distribution curve shows there is an increase in DO concentrations in the natural condition scenario, specifically in DO concentration values less than 5 mg/L in the existing condition run.

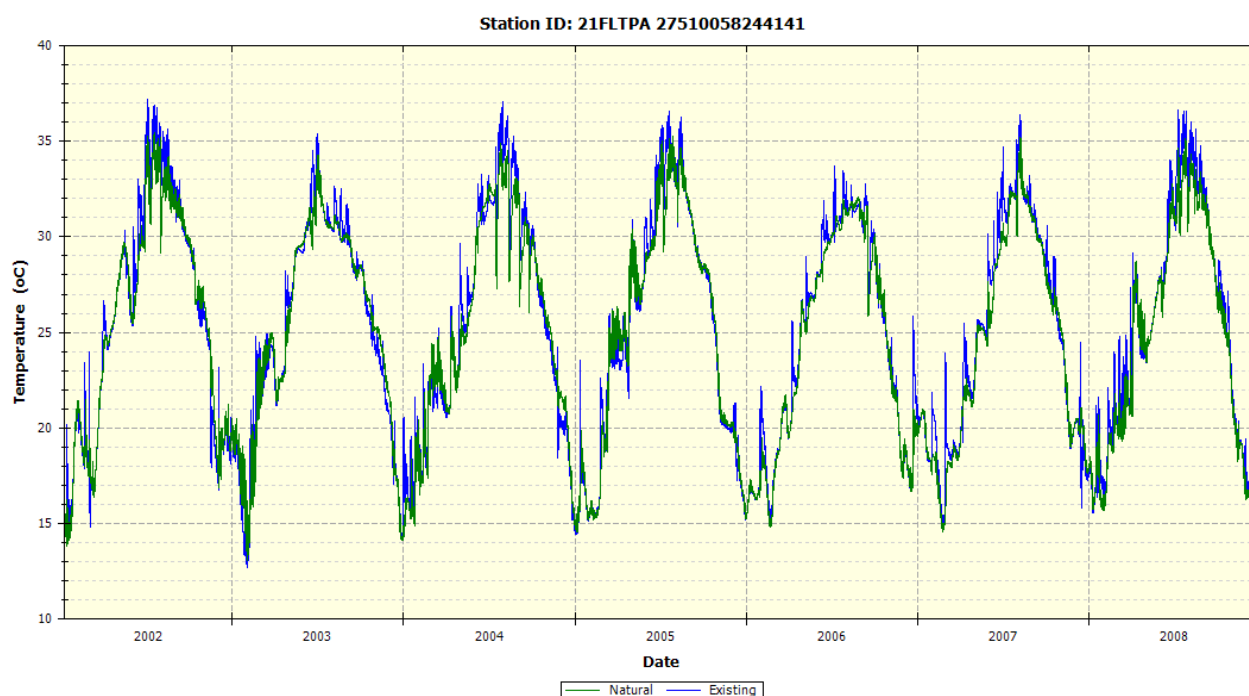


Figure 7.15 Existing condition temperature (°) verse natural condition temperature in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

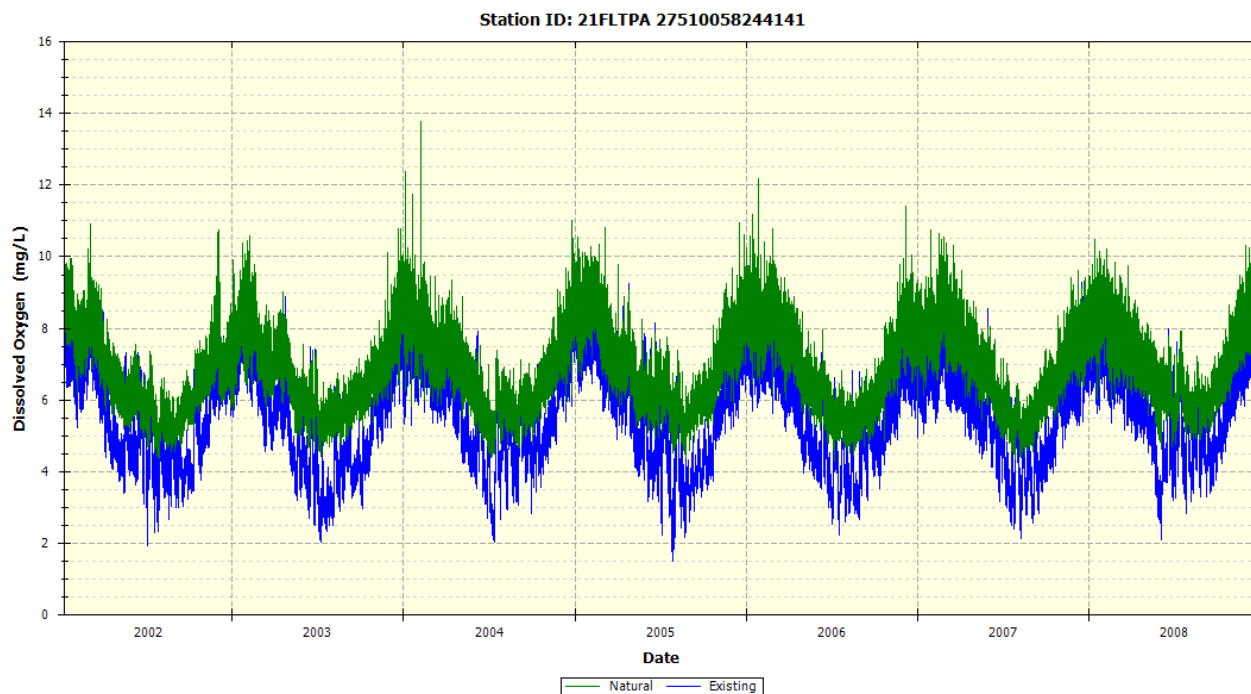


Figure 7.16 Existing condition oxygen (mg/L) verse natural condition dissolved oxygen in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

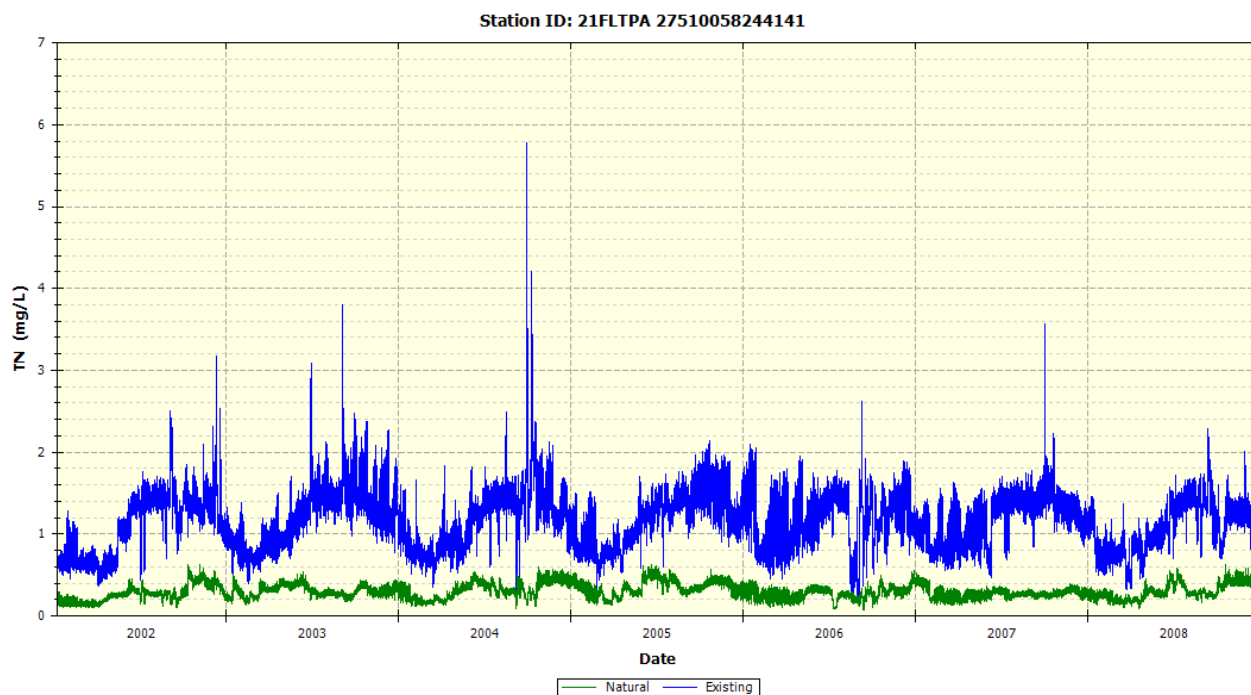


Figure 7.17 Existing condition total nitrogen (mg/L) verse natural condition total nitrogen in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

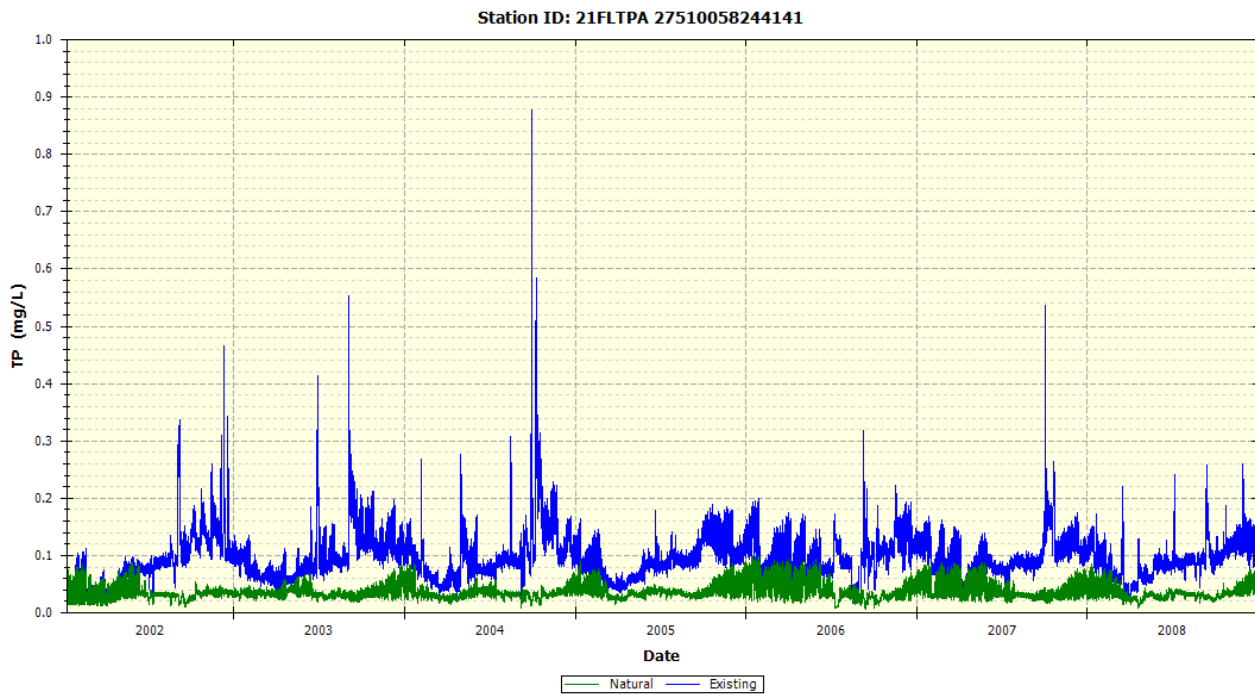


Figure 7.18 Existing condition total phosphorus (mg/L) verse natural condition total phosphorus in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

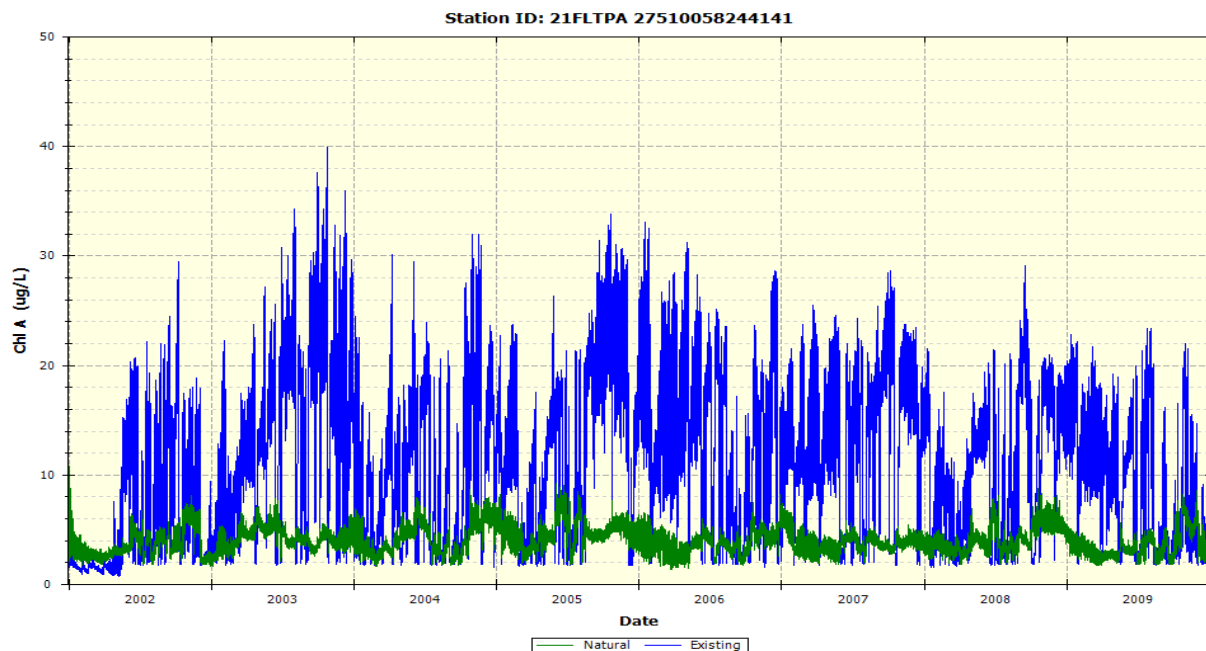


Table 7.19 Existing condition chlorophyll a (ug/L) verse natural condition chlorophyll a in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

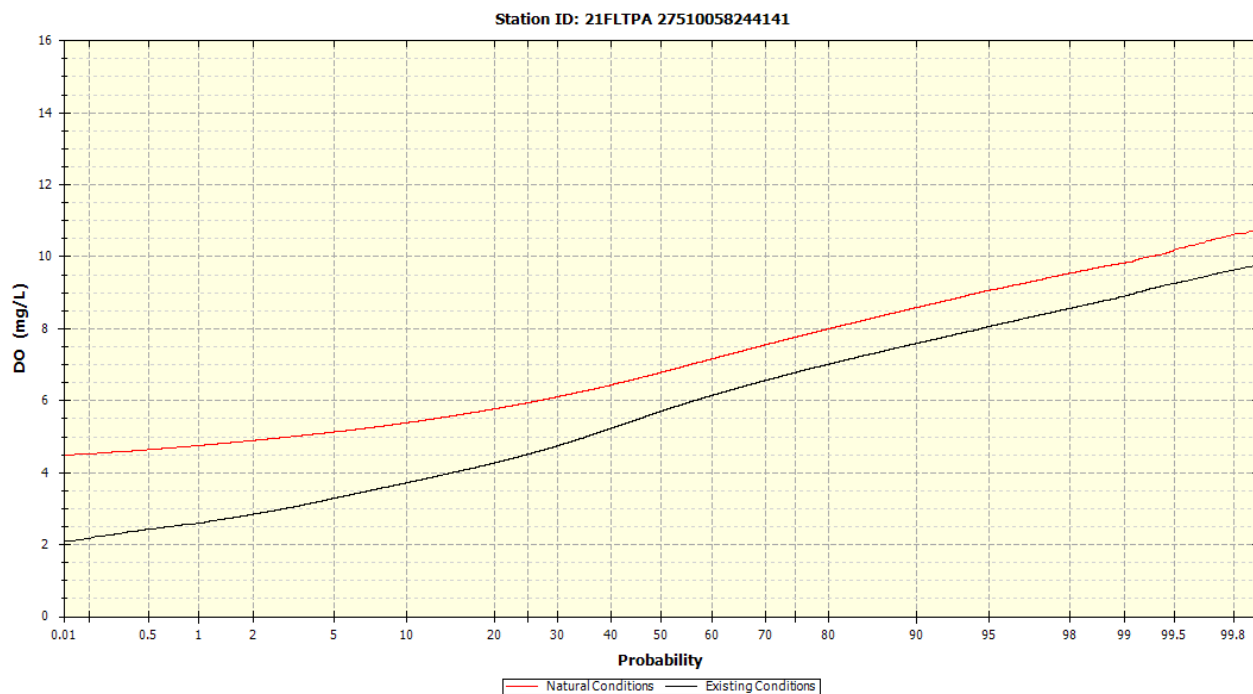


Figure 7.20 Dissolved oxygen concentration cumulative distribution function in Pinellas Park Ditch No. 1 at station 21FLTPA 27510058244141.

Table 7.3 Natural condition concentrations in the impaired WBID in Pinellas Park Ditch basin.

Parameter	WBID 1662
Total nitrogen (mg/L)	0.29
Total phosphorus (mg/L)	0.04
CBOD (mg/L)	1.18
DO (mg/L)	6.90

Table 7.4 Natural condition loadings in the impaired WBID in Pinellas Park Ditch basin.

Parameter	WBID 1409	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	1,200
Total phosphorus (mg/L)	--	300
BOD5 (mg/L)	--	5,422

8.0 TMDL DETERMINATION

The TMDL for a given pollutant and waterbody is comprised of the sum of individual wasteload allocations (WLAs) for point sources, and load allocations (LAs) for both nonpoint sources and natural background levels. In addition, the TMDL must include a margin of safety (MOS), either implicitly or explicitly, to account for the uncertainty in the relationship between pollutant loads and the quality of the receiving waterbody. Conceptually, this definition is represented by the equation:

$$\text{TMDL} = \sum \text{WLAs} + \sum \text{LAs} + \text{MOS}$$

The TMDL is the total amount of pollutant that can be assimilated by the receiving waterbody and still achieve water quality standards and the waterbody's designated use. In this TMDL development, allowable concentrations from all pollutant sources that cumulatively amount to no more than the TMDL must be set and thereby provide the basis to establish water quality-based controls. These TMDLs are expressed as annual geometric mean concentrations, since the approach used to determine the TMDL targets relied on geometric means. The TMDLs targets were determined to be the conditions needed to restore and maintain a balanced aquatic system. Furthermore, it is important to consider nutrient loading over time, since nutrients can accumulate in waterbodies.

The TMDL was determined for the concentrations coming from the upstream watersheds that directly drain into the listed WBID. During the development of this TMDL, it was determined that the natural condition scenario (removal of all anthropogenic sources and landuses) did not meet the Florida standards for DO. The DO was greater during the natural condition run, and nutrient loadings from the natural condition scenario were therefore used to determine the TMDL in accordance with the Natural Conditions narrative rule. EPA believes that setting the TMDL condition to a natural condition protects both 47(a) and 47(b) of the Florida narrative nutrient standard. It assures that no man induced activities would have caused an imbalance of flora and fauna. Natural background levels are presumed to protect aquatic life. Florida's water quality standards do not allow the abatement of natural conditions, this TMDL represents the lowest level of nutrients that could be established. By using the natural conditions nutrient loadings for the TMDL, the nutrient reductions also ensure protection of the downstream estuaries. The allocations for the WBID for total nitrogen, total phosphorus, and biochemical oxygen demand are presented in Table 8.1.

Table 8.1 TMDL load allocations for Pinellas Park Ditch No. 1, WBID 1662

Constituent	Current Condition		TMDL Condition		Percent Reduction		
	WLA (kg/yr)	LA (kg/yr)	WLA (kg/yr)	LA (kg/yr)	WLA	LA	MS4
Total Nitrogen	--	5,905	--	1,200	--	80%	80%
Total Phosphorus	--	697	--	300	--	85%	85%

Biochemical Oxygen Demand	--	16,620	--	5,422	--	67%	67%
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8.1 Critical Conditions and Seasonal Variation

EPA regulations at 40 CFR 130.7(c)(1) require TMDLs to take into account critical conditions for stream flow, loading, and water quality parameters. The critical condition is the combination of environmental factors creating the "worst case" scenario of water quality conditions in the waterbody. By achieving the water quality standards at critical conditions, it is expected that water quality standards should be achieved during all other times. Seasonal variation must also be considered to ensure that water quality standards will be met during all seasons of the year, and that the TMDL accounts for any seasonal change in flow or pollutant discharges, and any applicable water quality criteria or designated uses (such as swimming) that are expressed on a seasonal basis.

The critical condition for nonpoint source concentration and wet weather point source concentrations is typically an extended dry period followed by a rainfall runoff event. During the dry weather period, nutrients build up on the land surface, and are washed off by rainfall. The critical condition for continuous point source concentrations typically occurs during periods of low stream flow when dilution is minimized. Although loading of nonpoint source pollutants contributing to a nutrient impairment may occur during a runoff event, the expression of that nutrient impairment is more likely to occur during warmer months, and at times when the waterbody is poorly flushed.

8.2 Margin of Safety

The Margin of Safety accounts for uncertainty in the relationship between a pollutant load and the resultant condition of the waterbody. There are two methods for incorporating an MOS into TMDLs (USEPA 1991):

- Implicitly incorporate the MOS using conservative model assumptions to develop allocations
- Explicitly specify a portion of the total TMDL as the MOS and use the remainder for Allocations

This TMDL uses an implicit MOS since the TMDL targets for nutrients were set to natural background conditions.

8.3 Waste Load Allocations

Only MS4s and NPDES facilities discharging directly into lake segments (or upstream tributaries of those segments) are assigned a WLA. The WLAs, if applicable, are expressed separately for continuous discharge facilities (e.g., WWTPs) and MS4 areas, as the former discharges during all weather conditions whereas the later discharges in response to storm events.

8.3.1 Wastewater/Industrial Permitted Facilities

A TMDL wasteload allocation (WLA) is given to wastewater and industrial NPDES-permitted facilities discharging to surface waters within an impaired watershed. There are no continuously discharging NPDES-permitted point sources in WBID 1662, therefore no WLA was calculated.

8.3.2 Municipal Separate Storm Sewer System Permits

The WLA for MS4s are expressed in terms of percent reductions equivalent to the reductions required for nonpoint sources. Given the available data, it is not possible to estimate concentrations

coming exclusively from the MS4 areas. Although the aggregate concentration allocations for stormwater discharges are expressed in numeric form, i.e., percent reduction, based on the information available today, it is infeasible to calculate numeric WLAs for individual stormwater outfalls because discharges from these sources can be highly intermittent, are usually characterized by very high flows occurring over relatively short time intervals, and carry a variety of pollutants whose nature and extent varies according to geography and local land use. For example, municipal sources such as those covered by this TMDL often include numerous individual outfalls spread over large areas. Water quality impacts, in turn, also depend on a wide range of factors, including the magnitude and duration of rainfall events, the time period between events, soil conditions, fraction of land that is impervious to rainfall, other land use activities, and the ratio of stormwater discharge to receiving water flow.

This TMDL assumes for the reasons stated above that it is infeasible to calculate numeric water quality-based effluent limitations for stormwater discharges. Therefore, in the absence of information presented to the permitting authority showing otherwise, this TMDL assumes that water quality-based effluent limitations for stormwater sources of nutrients derived from this TMDL can be expressed in narrative form (e.g., as best management practices), provided that: (1) the permitting authority explains in the permit fact sheet the reasons it expects the chosen BMPs to achieve the aggregate wasteload allocation for these stormwater discharges; and (2) the state will perform ambient water quality monitoring for nutrients for the purpose of determining whether the BMPs in fact are achieving such aggregate wasteload allocation.

All Phase 1 MS4 permits issued in Florida include a re-opener clause allowing permit revisions for implementing TMDLs once they are formally adopted by rule. Florida may designate an area as a regulated Phase II MS4 in accordance with Rule 62-620.800, FAC. Florida's Phase II MS4 Generic Permit has a "self-implementing" provision that requires MS4 permittees to update their stormwater management program as needed to meet their TMDL allocations once those TMDLs are adopted. Permitted MS4s will be responsible for reducing only the loads associated with stormwater outfalls which it owns, manages, or otherwise has responsible control. MS4s are not responsible for reducing other nonpoint source loads within its jurisdiction. All future MS4s permitted in the area are automatically prescribed a WLA equivalent to the percent reduction assigned to the LA. The MS4 service areas described in Section 6.1.2 of this report are required to meet the percent reduction prescribed in Table 8.1, through the implementation of BMPs.

8.4 Load Allocations

The load allocation for nonpoint sources was assigned a percent reduction in nutrient concentrations from the current concentrations coming into WBID 1662 addressed in the TMDL report.

9.0 RECOMMENDATIONS/IMPLEMENTATION

The initial step in implementing a TMDL is to more specifically locate pollutant source(s) in the watershed. FDEP employs the Basin Management Action Plan (B-MAP) as the mechanism for developing strategies to accomplish the specified load reductions. Components of a B-MAP are:

- Allocations among stakeholders
- Listing of specific activities to achieve reductions
- Project initiation and completion timeliness

- Identification of funding opportunities
- Agreements
- Local ordinances
- Local water quality standards and permits
- Follow-up monitoring

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